

## Comparative utilization of phosphorus from sedimentary and igneous phosphate rock by major biotic components of aquatic ecosystem

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Received 30 April 2006; revised 28 November 2006; accepted 5 Desember 2006; available online 1 January 2007

**ABSTRACT:** Two laboratory experiments were conducted to evaluate the amount of P utilization from two different types of sparingly soluble phosphate rock by aquatic biotic communities. The first type was Mussoorie Phosphate Rock or MPR (sedimentary in origin) and other was Purulia Phosphate Rock or PPR (igneous in origin). The two trials were with eight different treatment combinations. Among various treatments, fish and Chironomid larvae contributed to some extent in increasing the available sediment phosphate content which in turn increased the soluble reactive phosphate (SRP) of overlying water. Concentration of SRP of overlying water decreased in the treatment with zooplanktons. Depletion of SRP of overlying water due to uptake of orthophosphate by *Chlorella* was also observed. The sedimentary type phosphate rock proved to be more efficient in releasing phosphate than igneous one.

**Key words:** Phosphate rock, putilization, fish, chironomid larvae, chlorella, zooplanktons

### INTRODUCTION

Phosphorus is the most studied element in aquatic system primarily due to the fact that it is the most limiting factor for primary productivity in water bodies and is essential for living organisms and is not exchangeable with other elements in biological system. It is an important constituent of biological systems and is a macronutrient but its availability is often extremely low (Hupfer, *et al.*, 2004). The effects of Phosphorus in nature are therefore profound. Usually phosphorus occurs in oxidized state, either as ions of inorganic orthophosphate or in organic compounds. Phosphorus in solution is normally considered to be orthophosphate (Soluble Reactive Phosphate or SRP) and is taken by different component members of an aquatic ecosystem. Boyd and Musig (1981) demonstrated that planktons in fish ponds absorbed an average of 41% of 0.30 mg/L addition of orthophosphate within 24 hours; however phosphorus that is not absorbed by planktons is rapidly absorbed mud (Hupfer, *et al.*, 2004; Cade-menum, 2005). Pelagic invertebrates not only transform, they can also translocate the recycled phosphorus within the system (Shapiro, 1984). Phosphorus is undoubtedly a recognized nutrient for pond fertilization. Among various phosphatic fertilizers

naturally occurring phosphate source (Mussoorie Phosphate Rock= MPR and Purulia Phosphate rock= PPR) can easily be used in fish farming ponds and has been amply proved to be an effective fertilizer in carp culture (Chakrabarty, 2006). The problem is its extremely low solubility in water. An attempt has been made here to quantify the difference of phosphorus utilization performances by various biotic members in laboratory condition from a marine sedimentary (MPR) and igneous (PPR) phosphate rock of Indian origin.

### MATERIALS AND METHODS

The chemical composition of MPR and PPR are as follows. The MPR is younger in geologic age than PPR (Table 1) and used as a cheap source of direct application fertilizer in fish farming ponds (Chakrabarty, 2006). In order to determine the extent of P-utilization from MPR and PPR by the major biotic components of the pond ecosystem, two laboratory experiments were carried out. The first experiment (trial-I) was carried out in 31 glass jars in the laboratory in presence of water and MPR all throughout, whereas the second one with water and PPR all throughout in the same manner with the previous one. For the first trial, all the 32 glass jars were filled with ground water and then subjected to following eight treatments in quadruplicate.

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Table1: Chemical data and characteristics of the two apatite sources (MPR and PPR) tested for their comparative efficiency in releasing phosphorus in laboratory condition

Major constituents %	MPR	PPR
P <sub>2</sub> O <sub>5</sub>	21.2	21.6
CaO	38.5	39.4
Fe	4.0	9.2
MgO	5.6	5.8
S	4.0	9.5
SiO <sub>2</sub>	6.6	8.1
C (Organic)	1.14	0.92
F	2.0	4.1
Characteristics	Mussoorie Phosphate Rock	Purulia Phosphate Rock
Unit cell dimension, a- axis, Å <sup>0</sup>	9.367	9.371
Mole ratio ( CO <sub>3</sub> :PO <sub>4</sub> )	0.052	0.013
Origin and nature	Marine sedimentary and carbonate apatite	Igneous and fluorapatite

Analysed at International Fertilizer Development Corporation, Alabama, USA.

(i) Water – (A<sub>1</sub>), (ii) Water + MPR – (B<sub>1</sub>), (iii) Water + Soil – (C<sub>1</sub>), (iv) Water + Soil + MPR – (D<sub>1</sub>), (v) Water + Soil + MPR + *Chlorella* sp. – (E<sub>1</sub>), (vi) Water + Soil + MPR + Zooplankton – (F<sub>1</sub>), (vii) Water + Soil + MPR + Chironomid larvae – (G<sub>1</sub>), (viii) Water + Soil + MPR + Fish – (H<sub>1</sub>).

For the second trial, all the 32 glass jars were filled with ground water and then subjected to above eight treatments in quadruplicate in presence of PPR in the same manner with the previous one.

(i) Water – (A<sub>2</sub>), (ii) Water + PPR – (B<sub>2</sub>), (iii) Water + Soil – (C<sub>2</sub>), (iv) Water + Soil + PPR – (D<sub>2</sub>), (v) Water + Soil + PPR + *Chlorella* sp. – (E<sub>2</sub>), (vi) Water + Soil + PPR + Zooplankton – (F<sub>2</sub>), (vii) Water + Soil + PPR + Chironomid larvae – (G<sub>2</sub>), (viii) Water + Soil + PPR + Fish – (H<sub>2</sub>).

Twenty gram of hundred mesh size ground MPR (trial-I) and PPR (trial-II) was placed in each of the experimental jars containing 2.5 l of ground water (pH 7.2). *Chlorella* sp. used in this experiment was procured from the laboratory axenic monoalgal culture (Chu– 10 medium). When the *Chlorella* concentration attained about 40 mg/L of Chl- a; 10 mL of such concentrate was dispensed in each experimental jar of treatment combination of E<sub>1</sub> and E<sub>2</sub>. *Daphnia* sp. was collected from the culture tank and concentrated to 150 numbers and used in each jar. Chironomid larvae (average length 0.5 cm ± 0.025) were locally procured and then acclimatized in the laboratory prior to their use in experiment (F<sub>1</sub> and F<sub>2</sub>). Thirty Chironomid larvae were used in each glass jar of treatment combination of G<sub>1</sub> and G<sub>2</sub>. Advanced fry of *Oreochromis mossambicus* (4.20 g ± 0.25 g; 4.0 cm ± 0.60) were procured and acclimatized well in the laboratory prior to their use.

Three acclimatized fry was then placed in each glass jar in treatment combination of H<sub>1</sub> and H<sub>2</sub>. Each set of glass jar was sacrificed at 0, 7 and 14 day for examination phosphorous contents of water, soil, *Chlorella* sp. All of them were carefully isolated, dried in hot air oven, grinded and total P content was measured using the method described by Jackson (1967). Chlorophyll-a (Chl- a) concentration of *Chlorella* sp. was analyzed following the method described by Vollenweider (1974). Orthophosphate concentration of water was also measured following standard methods (APHA, 2002). The results obtained in this study were statistically evaluated. Kruskal-Wallis one way analysis of variance by ranks was applied to find out the significance of difference among various treatments. Duncan's Multiple Range Test (Duncan, 1955) was also performed to test the significance of difference between every possible pair of treatment of trial-I and trial-II. This experiments were carried out in Dept. of Zoology, University of Kalyani (August 1993) as well as in Krishnagar Government College (August 2004). The results expressed here as mean of two experiments.

## RESULTS

Introduction of fish (H<sub>1</sub> and H<sub>2</sub>) and chironomid larvae (G<sub>1</sub> and G<sub>2</sub>) resulted in considerable rise (Figs. 1 and 2) of SRP concentration of water (0.1825 and 0.1425 mg/L) and sediments (13.0-12.9 and 11.7-11.5 mg kg/L) over other treatments (Figs. 3 and 4). Presence of zooplankton (F<sub>1</sub> and F<sub>2</sub>) in the MPR and PPR treatment, on the other hand, caused decline of SRP level of water all throughout. For example, the amount of orthophosphate (SRP) observed on day 0 (0.165 mg/L) declined to 0.0985 mg/L in (F<sub>1</sub>) on day 7 and finally to 0.065 mg/L on day 14.

Table 2: Results of Kruskal-Wallis one way analysis by ranks (H) and Duncan's multiple range test for mean values of SRP concentration of water in trial-I

Values of H (df=7)	Syi (df=7)	Duncan's multiple range test															
		R2	R3	R4	R5	R6	R7	R8		Comparison							
0 day 10.33 <sup>NS</sup>	1.154																
7 day 10.33 <sup>a</sup>	0.520	1.69	1.76	1.80	1.83	1.84	1.85	1.85	A	C	D	E	F	B	G	H	
										2.5	2.5<6.25	7.0<	9.25<11.25	14.0	14.5		
14 day 10.33 <sup>a</sup>	0.351	1.47	1.19	1.22	1.23	1.24	1.25	1.25	C	A	D	E	F	B	G	H	
										2.25	3.5	< 5	< 7.25	< 9.5	< 11.5	< 13.75	< 15.25

NS = Not significant, 'a' = significant at 5%

Horizontal bar indicates no significant differences

Table 3: Results of Kruskal-Wallis one way analysis by ranks (H) and Duncan's multiple range test for mean values of SRP concentration of water in trial-II

Values of H (df=7)	Syi (df=7)	Duncan's multiple range test															
		R2	R3	R4	R5	R6	R7	R8		Comparison							
0 day 10.33 <sup>NS</sup>	1.154																
7 day 10.33 <sup>a</sup>	0.526	1.67	1.74	1.78	1.81	1.82	1.83	1.85	A	C	D	E	F	B	G	H	
										2.5	2.5<6.25	7.0<	9.25<11.25	14.0	14.25		
14 day 10.33 <sup>a</sup>	0.357	1.45	1.17	1.20	1.20	1.21	1.21	1.21	C	A	D	E	F	B	G	H	
										2.2	3.5	< 5	< 7.2	< 9.5	< 11.5	< 13.25	< 14.25

NS = Not significant, 'a' = significant at 5%

Horizontal bar indicates no significant differences

In a same trend the amount of orthophosphate (SRP) observed on day 0 (0.125 mg/L) declined to 0.0985 mg/L in (F<sub>1</sub>) on day 7; and finally to 0.065 mg/L on day 14 in the glass jars with F<sub>2</sub> combination. The result was also true for *Chlorella* sp. the amount of orthophosphate of water in the MPR treatment with *Chlorella* sp. (E<sub>1</sub>) declined to 0.07 mg/L 0.05 on day 7 and further to 0.045 mg/L on day 14, from the initial concentration of 0.165. Similar result was observed with the same treatment combination (F<sub>2</sub>) with PPR (Fig. 2). Glass jars with only water (A<sub>1</sub> and A<sub>2</sub>) showed lowest SRP concentration (0.025 mg/L) of all, all throughout, whereas a stable and relatively high concentration of 0.155 to 0.1425 mg/L and 0.125 to 0.10 of SRP was maintained in the treatment combination of B<sub>1</sub> and B<sub>2</sub>. Kruskal-Wallis one way analysis of variance by ranks no significant difference (H=10.33; P>0.05) of treatment means in the beginning. Statistical analysis, however, showed significant difference (H=14.16; P<0.05) among treatments on day 7 and 14. Six treatment groups such as G<sub>1</sub> and G<sub>2</sub>; H<sub>1</sub> and H<sub>2</sub>; D<sub>1</sub> and D<sub>2</sub>; E<sub>1</sub> and E<sub>2</sub>, A<sub>1</sub> and

A<sub>2</sub>; C<sub>1</sub> and C<sub>2</sub> did not differ from each other on day 7 (Tables 2 and 3), whereas all the treatment groups except C<sub>1</sub> and C<sub>2</sub>; A<sub>1</sub> and A<sub>2</sub> showed significant difference on day 14 in both the trial. Significant difference was also found between H<sub>1</sub> and H<sub>2</sub> as well as G<sub>1</sub> and G<sub>2</sub> in respect of concentration of water SRP and sediment available -P (ANOVA, P< 0.05) on 0, 7 and 14 days of observation. The treatment combination with MPR always showed higher dissolution of P from insoluble P than the combination with PPR. The amount of Chl-a concentration (Fig. 5) did not varied significantly (ANOVA, P> 0.05) between the two treatment combination (E<sub>1</sub> and E<sub>2</sub>) throughout the experimental period. However, the total amount of phosphorus in chlorella (cell-P) was 0.05±0.002 mg Jar<sup>-1</sup> in day 0 of both trial-I and- II. The amount increased to 0.08±0.002 mg/Jar (trial-I) and to 0.078±0.003 mg/Jar (trial-II) on day 7 but declined to 0.065±0.002 mg/Jar (trial-I) and to 0.062±0.003 mg/Jar (trial II) in day 14. The cell-P also did not varied significantly in any day between two series.

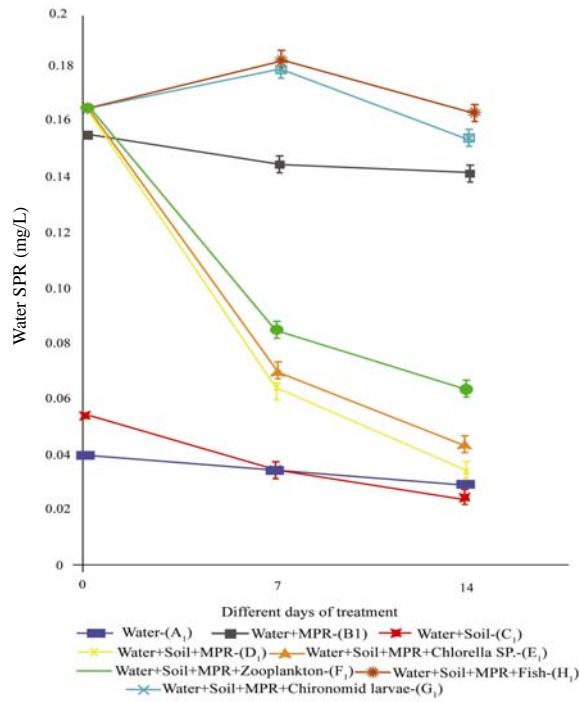


Fig. 1: Mean  $\pm$ SD concentration of soluble reactive phosphate of water in different days of various treatment of trial (H<sub>1</sub> and G<sub>1</sub>)

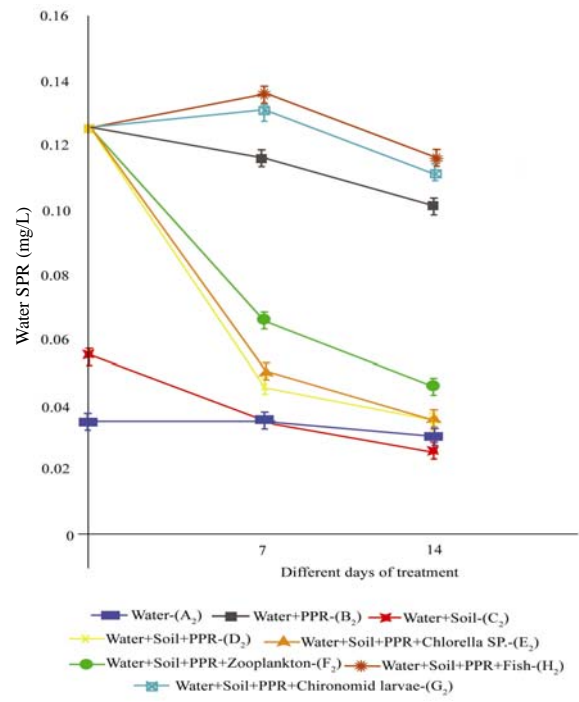


Fig. 2: Mean  $\pm$ SD concentration of soluble reactive phosphate of water in different days of various treatment of trial (H<sub>2</sub> and G<sub>2</sub>)

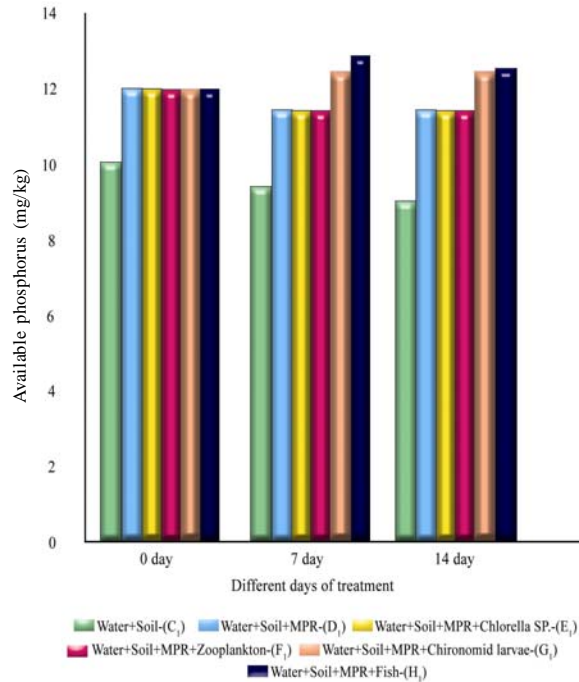


Fig. 3: Mean  $\pm$ SD concentration of soluble reactive phosphate of water in sediment of various treatment of trial (H<sub>1</sub> and G<sub>1</sub>)

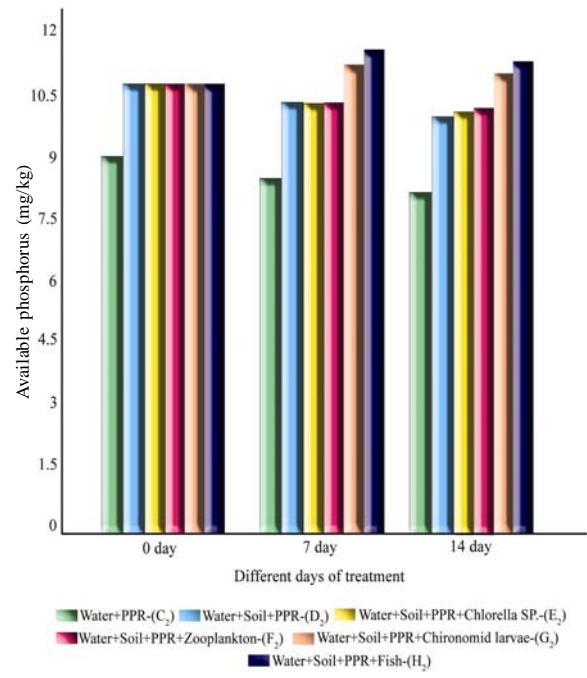


Fig. 4: Mean  $\pm$ SD concentration of soluble reactive phosphate of water in sediment of various treatment of trial (H<sub>2</sub> and G<sub>2</sub>)

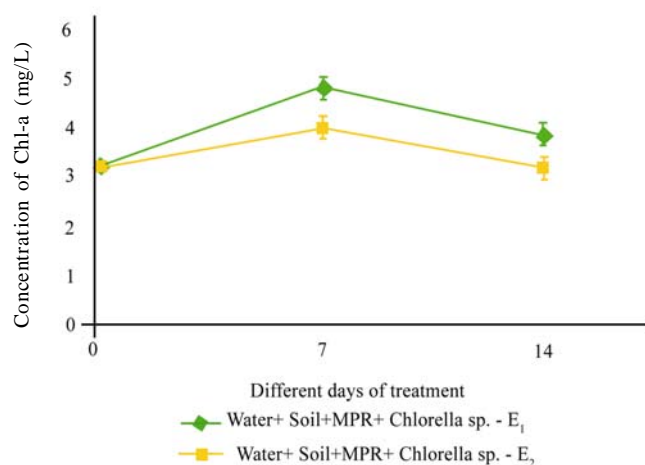


Fig. 5: Mean  $\pm$ SD concentration of chlorophyll-a in different days of trial-I ( $E_1$ ) and trial-II ( $E_2$ )

## DISCUSSION AND CONCLUSION

From the critical examination of the data it is revealed that fish and chironomid larvae contributed to some extent in increasing the available phosphate content of sediment which, in turn, increased the orthophosphate level of overlying water. This was due to their physical disturbance of the bottom sediments which induced the phosphorous release from sediment to the overlying water. The effect was perhaps brought about by the physical disturbance (Petr, 1977) and agitation of sediment enriched with phosphate rock. Similar results were obtained by Gabet, *et al.*, (2003), Schauser, *et al.*, (2003) and Sodergaard *et al.*, (2003). Biomanipulation trials clearly revealed that fish and chironomid larvae had a profound influence in the release of phosphate from otherwise insignificantly soluble phosphate rock by bioturbation (Chakrabarty, 2006). There was an increase in biomass of *Chlorella* evident from the rise in Chl-a content. The depletion in orthophosphate level of water ( $0.165-0.07 \text{ mg l}^{-1}$ ) was due to the phosphorous uptake by the *Chlorella*. This was evident from the increase in cell-P of *Chlorella* over time. Phosphorous was found to be a growth regulatory factor of algae by Ahlgren (1988). The critical examination of difference in liberation of available phosphorus from the two sparingly soluble phosphate sources indicated a significant difference ( $P < 0.05$ ) in treatment combination of  $H_1 - H_2$  and  $G_1 - G_2$ . This was because of the lesser unit cell dimension ( $> .008 \text{ \AA}$ ) of MPR than PPR (Table 1) as well as the higher  $\text{CO}_3:\text{PO}_4$  mole ratio of (0.052) of MPR than PPR (0.013). As lesser

unit cell dimension and higher  $\text{CO}_3:\text{PO}_4$  mole ratio helps in natural dissolution of phosphate rock in natural condition (PPCL 1987). The utilization of phosphorus by *chlorella* indicated no significant difference between the treatment combinations of  $F_1$  and  $F_2$ . Possibly, the amount of SRP was sufficient in both the treatments for algal growth. The concentration of cell P was also similar between two series. The study clearly indicated that MPR is certainly better phosphate fertilizer than PPR in terms of releasing phosphorus in aquaculture. It also proved that sedimentary phosphate rock has better applicability than igneous type. X-ray diffraction studies identified MPR as carbonate apatite and PPR as fluorapatite. Carbonate apatite is more responsive to natural dissolution than fluorapatite (PPCL, 1987). However, both the fertilizer can be used as direct application fertilizer in fish farming ponds with bottom grazing fishes. These fertilizers are environment friendly and cheap for fish culturist.

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**This article should be referenced as follows:**

*Chakrabarty, D., (2007). Comparative utilization of phosphorus from sedimentary and igneous phosphate rock by major biotic components of aquatic ecosystem. Int. J. Environ. Sci. Tech., 4 (1), 43-48.*