

Investigation on potential groundwater impacts and influence of local hydrogeology on natural attenuation of leachate at a municipal landfill

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ABSTRACT: The influence of local hydrogeology on natural attenuation of contaminants from landfill leachates in shallow aquifer underlying the active Olusosun landfill base in Lagos was investigated. In addition, the level of groundwater contamination in the vicinity of the landfill and of leachate migration pattern in groundwater down gradient of the landfill base was equally assessed. Landfill leachate and groundwater samples were collected and analyzed and characterized. Physico-chemical analyses of sampled water followed standard analytical methods. Analytical results showed a measurable impact of leachate outflows on groundwater quality. Elevated levels of anions: nitrate, chloride and sulphate in the groundwater body and heavy metals: Cr₃, Cd and Cu, were detected at measurable levels in groundwater down gradient of the landfill location without any particular attenuation pattern. The migration pattern and dispersion of leachates down gradient, 750 m away from the landfill location are irregular and difficult to predict as depicted by levels of contaminants present in groundwater. The study highlighted the importance of soil stratigraphy beneath the landfill base as an important factor in the natural attenuation of leachate constituents in the groundwater body.

Key words: Landfill, natural attenuation, migration, leachate, hydrogeology

INTRODUCTION

Groundwater contamination is a major concern in landfill operations because of the pollutional effects of landfill leachates and its potential health risks (Lee, and Jones-Lee, 1993a; 1993b; Christensen, *et al.*, 2001; Stollenwerk and Colman, 2003). The greatest contamination threat to groundwater comes from the leachate generated from the fill material which most often contains toxic substances especially when wastes of industrial origin are landfilled. However, it has been widely reported that leachates from landfills for non-hazardous waste could as well contain complex organic compounds, chlorinated hydrocarbons and metals at concentrations which pose a threat to both surface and ground waters. Solvents and other synthetic organic chemicals constitute a significant hazard, being of environmental significance at very low concentrations and resistant to degradation. Moreover, they may be transformed in some cases into more hazardous compounds (Baedecker, *et al.*, 1979 and Christensen, *et al.*, 2001). Most landfill leachate has high levels of BOD, COD, ammonia, chloride, sodium,

potassium, hardness and boron. The conditions within a landfill often vary over time, from aerobic to anaerobic thus allowing different chemical reactions to take place. The leachate from landfills for non-hazardous waste could produce reducing conditions at landfill base thereby enhancing the percolation of iron and manganese solution from the underlying deposits. The chemical composition of leachate varies due to a number of different known factors as the age, type of waste, operational practices at the site and percolation rate through the fill to the groundwater. Heavy metals such as cadmium, arsenic, chromium have been reported at excessive levels in groundwater due to landfills operations (Lee, *et al.*, 1986; Stollenwerk and Coleman, 2003). Bagchi (1989) described typical contaminants present in landfill leachate. Similarly, Lee and Jones (1993b) described typical composition of municipal landfill leachate. From the various studies over 40 organic compounds had been identified in leachate contaminated groundwater. Wood and Curran (1998) reported an average arsenic concentration in leachate contaminated groundwater to be as high as 21 µg/L with a range of 3.0-50 µg/L. An average arsenic concentration

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in the leachate plume as high as 326 $\mu\text{g/L}$ (range 52-647 $\mu\text{g/L}$) had also been reported (US Geological Survey, 2004). The volume of leachate depends principally on the area of the landfill, the meteorological and hydrogeological factors and effectiveness of capping. It is essential that the volume of leachate generated be kept to a minimum and also ensures that the ingress of groundwater and surface water is minimized and controlled. The volume of leachate generated is therefore expected to be very high in humid regions with high rainfall, or high run off and shallow water table (Chapman, 1992). This could well describe Lagos coastal environment where the topography is flat and low lying. The geology and hydrogeology of any potential landfill site thus have a major bearing on the level of natural protection for groundwater from contamination by landfill leachate. Even though there has been increasing concerns with regards to the vulnerability of groundwater to landfill operations in Lagos, Nigeria, there is however paucity of information as to the actual level of risk. Hence, there exists, a potential public health and environmental problems from existing solid waste management practices in Lagos State, the most critical being high risk to both surface water and groundwater quality through leachates outflow and infiltration (Lavalin, 1992; Longe and Kehinde, 2005). The main objective of this study was to investigate the influence of local hydrogeology on natural attenuation of contaminants from landfill leachates in shallow aquifer underlying the active Olusosun landfill base in Lagos. As part of the above overall stated objective, the determination of the level of groundwater contamination in the vicinity of the

landfill and of leachate migration pattern in groundwater down gradient of the landfill base was assessed. The study was of a particular interest as groundwater is a major source for domestic and industrial water supply in the mega city (Longe, *et al.*, 1987; Coode Blizard, 1997).

MATERIALS AND METHODS

Study area

The Olusosun landfill which is the focus of the current study occupies approximately 42 ha of land in Oregun, Ikeja Local Government Area, Lagos State. The active landfill area is made up of 37 hectares (Fig. 1). The site had been used in the past as a burrow pit where sand was mined for road construction. Initial excavation depth of 7 to 12 m existed before tipping of waste commenced. The landfill was designed to receive 7,365,000 tonnes of solid waste during its operational lifespan of 10 y. This Fig. represents a yearly average tipping volume of 736,500 tonnes. Operational design of waste to cover ratio of 9:1 was chosen for the ten years duration (Lavalin, 1992). The landfill accepts, officially, non-hazardous solid wastes of domestic, market, commercial, industrial and institutional origins but in practice all types of wastes are co-disposed. In order to ensure the integrity of the landfill under a natural attenuation design concept, a depth to seasonally high water table of 6metres from the landfill base was maintained. This is to allow for attenuation of leachate in the unsaturated zone. Its operation and maintenance were to be governed by stringent regulatory standards, but this has not been the case in practice.

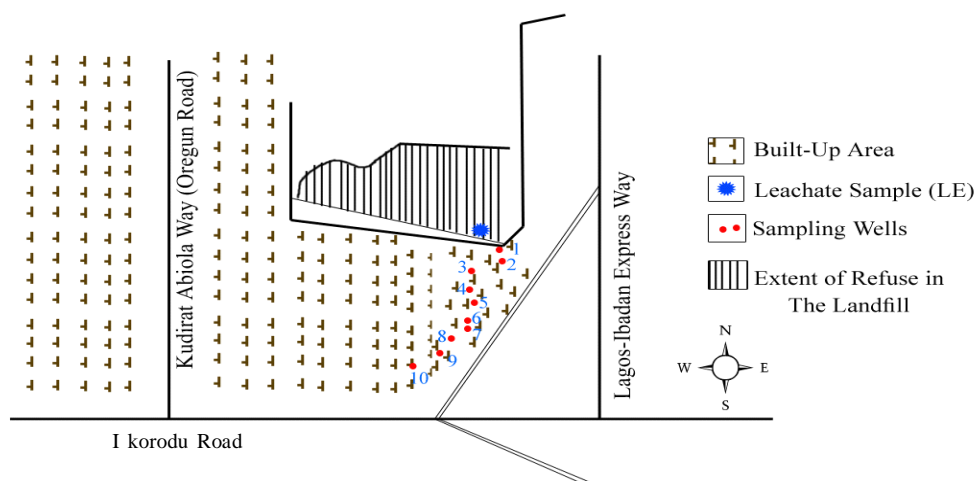


Fig. 1: A sketch diagram of the studied area with sampling points

Field sampling and laboratory analysis

Sampling points have been designated W1 to W10 representing ten wells previously installed approximately between 20 and 750 metres down gradient of the landfill location (Fig. 1). The ten water samples from the wells were used for this study. Groundwater sampling was carried out late July and early August 2004 during the raining season when a rise in water table was expected. A composite surface leachate sample scooped from the landfill base is designated LE. Both the leachate and groundwater samples were analyzed for parameters of interest characteristic to general mineral constituents in municipal landfill leachate such as total dissolved solids (TDS), chloride and heavy metals. The parameters were selected based on their relative importance in municipal landfill leachates composition, and their pollutional potential on groundwater resource in particular (Bagchi, 1989; Lee and Jones Lee, 1993b). Chloride was included in the groundwater quality assessment because of its measure of extent of dispersion of leachates in groundwater body (Chapman, 1992). All the sampled wells are protected sources of domestic water supply for private use. Each well was fitted with a surface pump. Groundwater samples were collected in plastic containers which were previously cleansed with non-ionic detergent and finally rinsed with deionized water prior to usage. As part of the quality control measures, samples were rinsed with sampled groundwater before being filled. Samples for the measure of heavy metals were collected in 1.5 L polyethylene bottles and preserved using 2 mL concentrated HNO₃ acid. All samples were preserved at 4°C and transported to the laboratory for analysis. Physico-chemical analyses of sampled water followed standard analytical methods. The pH was determined in-situ while the TDS was determined using a Horiba U-10 multiprobe meter. The dissolved oxygen level was measured using the Winkler's method. Levels of nitrate, phosphate, NH₃-N, and sulphate were determined using UV/ visible spectrophotometer, wavelength 200-1000nm, and chloride was determined by silver nitrate titration. Levels of heavy metals of interest, cadmium, copper, lead, chromium, iron, manganese and zinc were measured using Atomic Absorption Spectrometer with flame, graphite furnace and hydride system. The necessary hydrogeological study consisted of collation and analysis of existing hydrogeological data on strata logs, groundwater levels; and flow direction. Rainfall

data for over a period of 20 years from the nearest metrological station to the study area was collected and analyzed.

RESULTS

Geohydrological conditions

A major river, the Ogun is about 1km away north of the landfill location and it drains approximately 2,500 hectares of land. It flows in a southeasterly direction. The climate of the landfill location is consistent with the general climatic characteristics of the coastal environment in general and metropolitan Lagos in particular. There exists two principal climatic seasons, the rainy (wet) and dry seasons.

The wet season is from the month of April through October with a short break in mid-August, while the dry season is from the month of November through March. From the rainfall data analysis, the mean annual precipitation of the area is 1,526 mm which is lower than that of the entire Lagos metropolis of 3000 mm/year. This drier climate than what is generally obtainable in Lagos is expected to influence the volume of leachates production at the landfill site. Total monthly rainfall of greater than 175 mm characterized the months of May through Mid-August, while in June a monthly rate of 294 mm on average have been recorded. Fig. 2 presents mean precipitation from the landfill site for a ten year period. Recorded total monthly rainfall values for the months of September and October are 194 and 172 mm respectively with a decline as the dry season sets in gradually. Between the dry season months of November and March, recorded total monthly rainfall is low, values range between 28.5mm and 100.0 mm while a monthly average of less than 20 mm characterized the month of January. An estimated mean annual potential evapo-transpiration of the landfill site had been evaluated as 98 cm (Lavalin, 1999).

Hydrogeology

The hydrogeological condition of the landfill site is consistent with the regional hydrogeological setting of Lagos area as depicted by Longe, *et al.*, (1987). The sub-surface geology of the landfill reveals a lateritic cover, reddish brown in colour with sand and clay portions. The lateritic cover is of variable thickness averaging 4metres. This lateritic cover overlies an alternating sequence of sand and clay deposit (Fig. 3). Underlying this is a water-bearing zone consisting of loose, medium to coarse sand with an average thickness of 10.4 m.

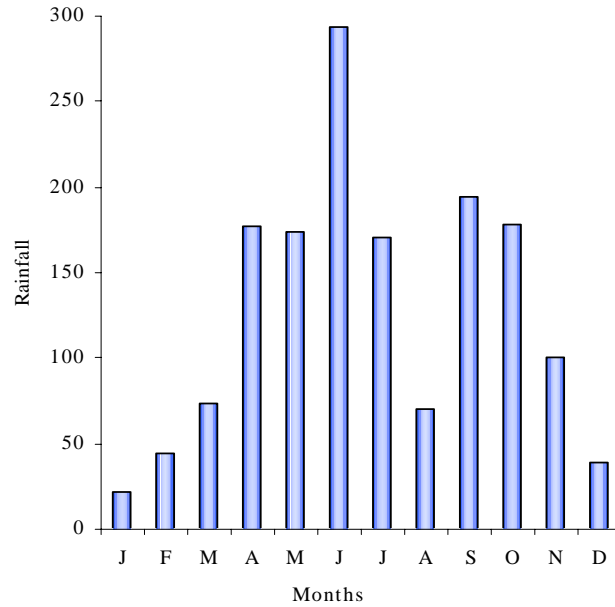


Fig. 2: Average rainfall in Ikeja area of Metropolitan Lagos

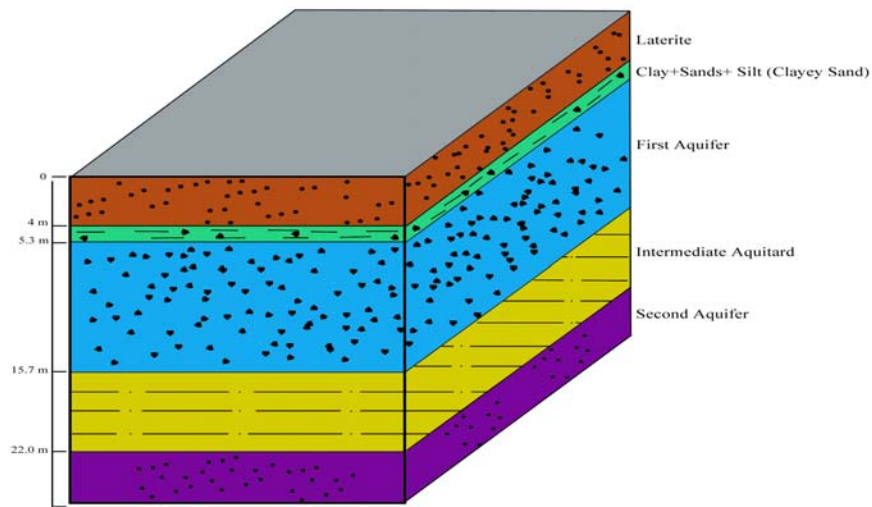


Fig. 3: A schematic block diagram of subsurface geology of the landfill location

Depths of occurrence vary from 4 to 7 m at the northern part of the landfill to 10 m at the southern end. This represents the first aquifer horizon described by Longe, *et al.*, (1987). The aquifer has a predominant lateral groundwater movement and it is prone to contamination. The calculated hydraulic conductivity of this water-bearing zone is in the order of 10^{-3} cm/s. This aquifer horizon is immediately underlain by a semi-permeable layer of reddish gray, silt with clay seal of variable thickness averaging at 4 m. Hazen estimates

from grain size analysis, indicates hydraulic conductivity of 10^{-7} cm/s for the semi-permeable layer that separates the first aquifer from the underlying aquifer horizon. The second aquifer consists of medium sand with traces of silt and exhibits hydraulic conductivities of the order of 10^{-4} to 10^{-3} cm/s. The first and second aquifer horizons are major sources of domestic and industrial water supply and both are the main thrust of this work.

Groundwater quality

Analytical results of physico-chemical characteristics of leachates and groundwater samples from wells are presented in Tables 1 to 4. The groundwater is generally acidic with pH values varying between 3.90 and 5.27, while the leachate is slightly neutral. The mean pH value of 4.66 recorded for the groundwater in this study is below the stipulated WHO pH tolerance range of 6.5 to 8.6 for

potable water (WHO, 2004). Water from W8 at 600 metres away from the landfill has the lowest pH of 3.9, hence the most acidic in this present study. The total dissolved solids (TDS) values obtained in groundwater varied from 201 to 800 mg/L, while the TDS value obtained for the leachate was 3,215 mg/L. A significant attenuation of TDS concentrations in all wells could be observed down gradient of the landfill site but without any particular pattern.

Table 1: Physico-chemical characteristics of leachate and groundwater

Sample	Distance (m)	pH	TDS (mg/L)	Ca (mg/L)	NH ₄ ⁺ (mg/L)	SO ₄ (mg/L)	NO ₃ (mg/L)	Cl ⁻ (mg/L)	PO ₄ ⁻ (mg/L)
LE	0	7.36	3215	1460	1.50	25	2.5	1000	3.5
W 1	20	4.50	241.5	105.9	0.15	25.15	0.12	75.47	0.21
W 2	50	5.03	259.0	237.3	0.16	26.76	0.21	80.50	0.04
W 3	400	4.22	201.0	168.0	0.16	20.88	0.21	62.66	0.01
W4	450	4.69	214.0	113.2	0.14	22.34	0.31	67.03	0.03
W5	500	5.27	289.5	262.8	0.18	30.15	0.12	90.46	0.01
W6	550	4.77	250.0	219.0	0.16	26.09	0.21	78.28	0.05
W7	560	4.06	292.0	109.5	0.19	30.42	0.32	91.25	0.03
W8	600	3.90	615.0	146.0	0.38	64.00	0.21	192.0	0.02
W 9	650	5.06	800.0	277.4	0.50	80.33	0.43	250.0	0.05
W10	750	5.14	272.5	292.0	0.17	28.39	0.22	85.16	0.03
WHO STD		6.5-8.5	500	200	0.5	400	10	250	5.0

Table 2: Descriptive statistics for physico-chemical parameters of leachate and groundwater

Parameter	Min	Max	Mean	SD	SE	Variance	Median	Range
pH	3.90	5.27	4.66	0.48	0.15	0.23	4.73	1.37
TDS	201.00	800.00	343.45	198.91	0.15	39564.08	265.75	800.00
Ca ⁺	105.90	292.00	193.11	73.18	23.14	5355.05	193.50	292.00
NH ₄ ⁺	0.14	0.50	0.22	0.12	0.038	0.01	0.16	0.36
SO ₄ ⁺	20.88	80.33	35.45	19.96	6.31	398.43	27.58	59.45
NO ₃ ⁻	0.12	0.43	0.24	0.09	0.03	0.01	0.12	0.31
Cl ⁻	62.66	250.00	107.28	62.14	19.65	3861.86	82.83	187.34
PO ₄ ⁺	0.01	0.21	0.05	0.06	0.02	0.003	0.03	0.20

Calcium concentrations in the groundwater varied between 105.9 and 292.0 mg/L. A mean value of 193.1 mg/L was obtained which fell below the WHO stipulated tolerance level of 200 mg/L for potable water. However water from five out of the ten wells tested: w₂, w₅, w₆, w₉ and w₁₀ had values far above the WHO stipulated tolerance level. The presence of calcium in water generally varies according to the proximity of natural sources. Sewage and industrial wastes are known to account for such anthropogenic sources of calcium in water. The measured values of ammonia, sulphate, nitrate and phosphate in leachate sample and in groundwater were below the WHO stipulated value for each of the anion for potable water (Table 2). The exception in this observation is the concentration of NH₄⁺; the leachate has an elevated value of 1.50 mg/L, while the

concentrations in the groundwater ranged from 0.14 to 0.50 mg/L. The presence of NH₄⁺ in groundwater has biochemical significance as a useful indicator of organic pollution (Chapman, 1992). Measured chloride concentration in leachate was 1000 mg/L while its concentration in groundwater ranged between 62.66 and 250.0 mg/L. There was no definite attenuation pattern observed in wells down gradient of the landfill site. Chloride is a good measure of the extent of the dispersion of leachates in groundwater. However, high chloride concentrations are yet common in groundwater in Lagos but, an increasing chloride concentrations in the coastal groundwater have been reported (Coode Blizard, 1997). The current observed chloride concentration in groundwater samples are above the background chloride levels in aquifer underlying Lagos metropolis.

The heavy metals

Measured values of heavy metals in landfill leachate and groundwater are presented in Table 3, while Table 4 presents the descriptive statistics of the heavy metals. The trace metals rarely occur in groundwater at concentrations large enough to comprise a significant percentage of the total dissolved solids. Except for iron, heavy metals in natural or contaminated groundwater almost invariably occur at concentrations below 1 mg/L (Chapman, 1992). The measured values of Fe, Mn and Pb⁺ in the landfill leachate were above the WHO tolerance levels for drinking water quality. High iron

Lead (Pb⁺) was not detected in any of the well. Its non detection in groundwater may not be unconnected with its affinity to be absorbed by clayey soils. Zinc and chromium were not detected in the landfill leachate but were found in appreciable levels in the groundwater. Measured values of Zn ranged from 0.01 to 0.40 mg/L, a mean value of 0.17 mg/L, which is far below the WHO tolerance value of 5.0 mg/L specified for potable water was obtained. However, high levels of chromium were measured in all wells. The mean value of 0.89 mg/L (range 0.02-1.71 mg/L) is far above the WHO acceptable level of 0.05 mg/L.

Table 3: Heavy metals concentrations in leachate and groundwater

Samples	Distance (m)	Fe (mg/L)	Mn (mg/L)	Cu (mg/L)	Zn (mg/L)	Pb (mg/L)	Cr ³ (mg/L)	Cd (mg/L)
LE	0	1.056	0.35	ND	ND	0.113	ND	ND
W1	20	0.41	0.08	0.35	1.75	ND	1.13	ND
W2	50	0.64	0.13	0.05	0.25	ND	0.30	ND
W3	400	0.52	0.10	0.01	0.05	ND	0.95	ND
W4	450	0.51	0.10	0.31	1.55	ND	0.74	ND
W5	500	1.11	0.22	0.40	2.00	ND	0.85	ND
W6	550	0.60	0.13	0.17	0.85	ND	0.20	ND
W7	570	0.40	0.08	0.08	0.40	ND	1.67	ND
W8	600	0.82	0.16	0.03	0.15	ND	0.34	ND
W9	650	4.19	0.84	0.35	1.75	ND	1.71	ND
W10	750	0.44	0.09	0.02	0.10	ND	1.02	ND
WHO		0.3	0.1	1.0	5.0	0.05	0.05	0.01

ND =Not detected

Table 4: Descriptive Statistics for heavy metals in groundwater

Parameter	Min	Max	Mean	SD	SE	Variance	Median	Range
Fe	0.4	4.19	0.96	1.15	0.37	1.33	0.56	3.79
Mn	0.08	0.84	0.19	0.23	0.07	0.05	0.11	0.76
Cu	0.01	0.40	0.17	0.16	0.05	0.02	0.12	0.39
Zn	0.01	0.40	0.17	0.16	0.05	0.02	0.12	0.39
Pb	0.00	0.00	0.00	0.00	0.16	0.02	0.12	0.39
Cr	0.20	1.71	0.89	0.53	0.16	0.28	0.90	1.51

concentrations were measured in groundwater with no discernable attenuation trend down gradient of the landfill site. However, high iron concentrations in groundwater have been widely reported in the country (WHO/UNICEF, 2006). Measured values of iron in groundwater ranged from 0.41 to 4.19 mg/L, the mean value of 0.94 exceeded the WHO tolerance level of 0.30 mg/L for potable water, thereby making iron an important groundwater quality issue in the study area. Concentrations of manganese in all the wells ranged between 0.08 and 0.84 mg/L with a mean value of 0.193 mg/L. Concentrations of manganese in excess of 0.2 mg/L makes water distasteful to drink with no specific toxic effects. An acceptable level of 0.05 mg/L set by WHO for drinking water quality was not met by the

groundwater quality. Heavy doses of chromium salts even though are rapidly eliminated from human body, could corrode the intestinal tract (WHO, 2004). Levels of copper (Cu²⁺) measured in groundwater ranged between 0.01 and 0.40 mg/L less than the WHO specified tolerance level of 1.0 mg/L for drinking water quality.

DISCUSSION AND CONCLUSION

The general acidic nature of groundwater is could not be solely attributed to leachate contamination. High iron level is characteristic of groundwater quality in Nigeria in general and in Lagos in particular (WHO/UNICEF, 2006). For the anions, there is no definite attenuation pattern observed in wells down gradient

of the landfill site. For instance, the observed TDS concentrations of 615 and 800 g/L obtained in wells W8 and W9 respectively at distances of 600 and 670 m corroborates this observation. Aside from these two wells, the TDS values of water from all the other wells fell below the WHO tolerance level of 500 mg/L for drinking water quality. Nearly all trace metals of interest in groundwater problems are influenced by redox conditions, as a result of either of changes in the oxidation state of the trace metal or non-metallic elements with which it forms complexes. The groundwater often acts as a redox environment due to changes in solid phases in the porous medium with adsorption of the trace metal. This may partly be responsible for the occurrence of trace metals at low concentrations in the groundwater samples. The presence of Pb^{+} in the landfill leachate and its absence in all the wells down gradient of the landfill is best explained by the presence of clayey soil. The mobility of Pb^{+} in clayey materials is low. In soil systems, Pb^{+} can also form poorly soluble precipitates with sulphate, carbonate, phosphate and sulphide anions. In a municipal landfill, where anaerobic conditions are likely to occur, Pb^{+} should become more mobile, but this is not so in this case. Even though the soil material that favours its attenuation in leachate includes organic matter, clays, and free lime, the only possible contributing factor here is clay. The sub-soil stratigraphy of the study area which is characterized with the presence of an unsaturated zone of about 6 m above the water table is expected to equally play a major role. The unsaturated zone which consists of silty clay could as well play a significant role in the natural attenuation of leachate in groundwater body. The observed attenuation pattern of contaminants in wells up to the distance of 750 m away from the landfill site is irregular and unpredictable. The local geology of the study area appears to play a major role in this respect. The noticeable stratigraphic variations depicting the local geology (Fig. 3) may have influenced natural attenuation of contaminants, their pattern of transfer and subsequent breakthrough into the groundwater body. Heterogeneities are of significant importance in the aquifers underlying Lagos metropolis (Longe, *et al.*, 1987), hydraulic conductivity contrasts do occur, due to intercalations of semi-permeable and impermeable layers. This existing hydrogeological setting favours adsorption and retention of contaminants in the pore spaces, thereby

making pattern of leachate dispersion very irregular and difficult to predict. As much as some of the contaminants are found in groundwater at concentrations higher than the background groundwater chemistry of aquifer in Lagos, there is therefore the tendency to suspect the landfill as one of the sources of groundwater contamination in the study area. The study has highlighted the importance of soil stratigraphy beneath the landfill base as an important factor in the attenuation of leachate constituents. The local hydrogeological setting might have contributed to the enhancement of the natural attenuation of contaminants to a certain degree. However, impairment of groundwater quality is not totally removed or eliminated. There is a great tendency for an increase in groundwater contamination around the landfill location considering the current levels of chloride, nitrate and organics found in the groundwater which are above background levels. In order to complement this preliminary result, further study on groundwater quality in and around the landfill in all directions other than in one as presently carried out will be necessary. This is to further ascertain the suitability or otherwise of the natural attenuation landfill type, and confirm the efficiency of the influence of the local sub-soil stratigraphy as well as the hydrogeology in groundwater protection from leachate outflows at the landfill site. This becomes very important as confirmatory result could promote natural attenuation type landfill as an alternative to the expensive containment and engineered landfill in developing nations of the world where financial resources are limited and technical or managerial capability to operate such facility are lacking. This could then guarantee a reduction in groundwater contamination resulting from unplanned municipal solid waste management. In addition to the above, in the interim, in order to protect the groundwater quality of the study area, a monitoring programme for groundwater quality status around the vicinity of the landfill is suggested. Monitoring wells should be sited both in the up gradient and down gradient of the landfill. Groundwater monitoring wells of 10 to 15 m intervals should be sited along the down gradient perimeters, southeasterly and western edges of the landfill. At least two number wells should be sited at approximately 5 to 10 m intervals along the up gradient in the perimeters, N-W edges of the landfill to monitor background water quality. Finally, for sitting new landfills, the potential for pollution of

groundwaters could be drastically minimized by maintaining an adequate buffer zone between the landfill and the property line of the adjacent property. A buffer zone of three miles in the direction of groundwater flow is most appropriate (Lee *et al.*, 2005).

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