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Kinetic mechanism of competitive adsorption of disperse dye and anionic dye on fly ash

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Abstract The adsorption behavior of four anionic dyes and one disperse dye in single solution and binary solutions on fly ash was investigated in order to elucidate the effect of competitive adsorption on the kinetics. The experimental findings showed that adsorption equilibriums of four anionic dyes were reached within 50 min either in the single solution or in binary mixtures. Competitive adsorption increased the time of attaining equilibrium of disperse dye. Desorption of dyes suggested the predominant adsorption mechanisms, that is, chemisorption for anionic dyes and physisorption for disperse dye. For the binary mixtures, the anionic dyes could be adsorbed preferentially on fly ash at the first stage. Second-order kinetic models fitted better to the equilibrium data of all dyes in the single solution as well as in the binary mixtures. The maximum rate constant of intraparticle diffusion and the minimum external mass transfer coefficient was found for disperse dye both in single and in binary solutions. The intraparticle diffusion constants and external mass transfer coefficients of the four anionic dyes in binary solution are similar to those obtained in single solution. The Biot number confirmed that the intraparticle diffusion was the rate-limiting step in the dye sorption process.

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Keywords Binary solutions · Desorption · External mass transfer · Second-order kinetic modal · Intraparticle diffusion · Wastewater

Introduction

Synthetic dyes have been increasingly used in textile, leather, food processing, dyeing, cosmetics, and so on. Wastewater streaming from these industries contains unused dyes (about 8-20 % of the total pollution load due to incomplete exhaustion of the dye) (Mukherjee et al. 1999), which causes adverse effects on aquatic environment. Because the dyes in low concentration give water undesirable color (Wong et al. 2003), they reduce sunlight penetration into water and then affect photosynthesis of aquatic organisms. Most dyes with a complex aromatic molecular structure are considered to be non-oxidizable substances by traditional biological and physical treatment (Mane et al. 2007). As a result, the conventional techniques of wastewater treatment on the basis of biological oxygen demand removal are largely ineffective in removing colored textile dyes from effluent. Adsorption process, therefore, is an attractive and effective alternative treatment for dye removal from wastewater.

Fly ash (FA) is a by-product produced during the combustion of coal in the electricity generation process. Although fly ash has also been used effectively for building materials, soil amendment, and fillers, it is unlikely that these applications will result in the reuse of all the fly ash generated. Great efforts are therefore being made to explore other applications prior to disposal. Since fly ash is enriched with SiO₂ and contains a portion of unburned carbon, this waste possesses a potentiality of a low-cost adsorbent to remove various hazardous materials from



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wastewater (Dizge et al. 2008; Hsu 2008; Janoš and Buchtová 2003; Kao et al. 2000; Mohan et al. 2002; Ramakrishna and Viraraghavan 1997; Wang et al. 2005; Wang and Wu 2006). Recently, the technical feasibilities were reviewed by Ahmaruzzaman (2010) and Saka et al. (2012).

Numerous investigations have been reported on the adsorption of single component dye by various adsorbents. However, due to complexity of the textile effluent and the variability of the dyeing process little successes have been reported in using this technique as a full-scale process to decolorize textile wastewater (Al-Degs et al. 2007; Nouri et al. 2002). Since many industrial effluents contain a mixtures of several dyes, the study of multi-component adsorption presents a major challenge compared to the single component adsorption. This work concerns the single and binary adsorption of non-ionic dyes (Disperse red 74) and anionic dye (Acid blue 193, Acid black 1, Reactive blue 4 and Reactive blue 171) by fly ash.

Materials and methods

Adsorbents and characterization

Raw coal fly ash used in this study was provided by a fullscale power plant in China, which was collected at the electrostatic precipitator. The chemical compositions of the sample were measured using a NK-09 inductively coupled plasma atomic emission spectroscopy (*Beijing Zhuochuan Electronic Technology*), which mainly contained SiO₂ (47.2 %), Al₂O₃ (24.8 %), Fe₂O₃ (15.9 %), and CaO (4.9 %) (obtained from the plant). The loss of ignition (LOI) was 3.81 %, which measured by a standard combustion method.

Raw fly ash dried at 105 °C overnight was separated manually by shaking stainless steel mesh screens with the opening of standard numbers of 160 and 200. Fly ash particles with size ranging from 75 to 96 μ m were pretreated in a furnace at 350 °C for 30 min, and kept in a desiccator for further study. The fly ash density was 1,838 kg m⁻³, which was determined by the displacement method. The mean particle size (85.5 μ m) was calculated by the average value of the two sizes of sieve opening. Prior to their use, the adsorbents were oven-dried at 105 °C for 2 h to eliminate traces of moisture.

The surface area and total pore volume of the FA were measured through N₂ adsorption at -196 °C using a TRISTAR-3000 surface area and porosity analyzer (*Micromeritics*). The specific surface area and pore volume of the FA was found to be 9.84 m² g⁻¹ and 0.023 cm³ g⁻¹, respectively. pH of zero charge (pH_{pzc} = 8.4) of the samples was determined using pH drift method by a

PHS-3C pH meter (Shanghai Precision & Scientific Instrument Co. LTD) (Fariaa et al. 2004).

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Adsorbates

The adsorbates in this study include one non-ionic dye, Disperse Red 74 (DR74), and four anionic dyes, Acid Black 1 (AB1), Acid Blue193 (AB193), Reactive blue 4 (RB4), and Reactive Blue 171 (RB171). The anionic dyes were supplied by Dystar and the disperse dye was purchased from Baodao Chemical Engineering Co. Ltd. All dyes were employed without further purification. Their stereo-chemical structures were optimized by ChemSketch 12.0 (Advanced Chemistry Development, Toronto, Canada) and displayed in Fig. 1. Molecular weight and dimension data were obtained using ChemSketch 12.0. Table 1 compared some selected chemical properties of these dyes.

The synthetic wastewaters were prepared by dissolving dyestuffs in de-ionized water to produce a stock solution of 1 g L⁻¹. The initial pH values of synthetic wastewaters were in the range 5.6–6.3, and the pH values of residue solution were kept in 5.8–7.4 in order to eliminate the effect of dye self-aggregate (Mall et al. 2005). For batch study, aqueous solutions of these dyes were prepared from stock solutions in de-ionized water. The UV–Visible spectrum of each dye solution was previously determined and identified the maximum absorption wavelength (λ_{max}).

Kinetic adsorption study

Batch experiments were carried out to measure the competitive adsorption characteristics between non-ionic dye and anionic dyes on FA. A predetermined amount of the fly ash sample was added to 100 mL of dye solution. The mixture was stirred with mechanical agitator (150 rpm) for various times to determined the equilibrium time at 25 °C. After stirring, the suspensions were separated by filtering. Dye solutions included five single component systems, DR74, RB4, RB171, AB1, and AB193, and four binary component systems, DR74/RB4, DR74/RB171, DR74/ AB1, and DR74/AB193. For the binary component systems, all solutions were prepared with solutions of equal mass concentrations. The initial concentration of each dye was kept 100 mg L^{-1} either in single solutions or in binary systems. All adsorption data reported in this paper were the average values of three runs.

The residual concentration of dye solution was determined using a UV-2102PC UV-visible spectrometer (*Unicol Instrument Co., Ltd*). For the single dye systems, concentrations were measured using a calibration curve prepared at the corresponding maximum wavelength. For the binary systems, the value of the unknown dye



Fig. 1 Stereo-chemical structure of dyes. a Disperse Red 74, b Reactive blue 4, c Reactive Blue 171, d Acid Blue193, e Acid Black 1







Table 1Chemical properties ofthe dyes	Property	DR74	RB4	RB171	AB1	AB193
	Molecular weight ^a	456.5	620.4	1,281.0	570.5	836.8
	Width (Å)	18.042	15.087	31.986	21.658	15.889
	Depth (Å)	9.436	10.605	14.086	6.192	13.734
	Thickness (Å)	3.042	2.593	6.207	4.751	3.644
	$\lambda_{\rm max}$ (nm)	522	600	602	614	592
	Number of ionizable group	0	2	6	2	2
^a Does not include associated sodium ions	Percent of ionizable group (%) ^a	0	15.47	22.95	16.82	11.47



concentrations could be determined by solving simultaneous linear equation 1 and 2 (Porter et al. 1999).

$$A_1 = \varepsilon_{11} b C_1 + \varepsilon_{21} b C_2 \tag{1}$$

$$A_2 = \varepsilon_{12} b C_1 + \varepsilon_{22} b C_2 \tag{2}$$

where A_1 and A_2 are the absorbance at the maximum wavelength of dye 1 and 2 ($\lambda_{max,1}$ and $\lambda_{max,2}$), respectively. ε_{11} and ε_{12} are the absorption coefficients of dye 1 at $\lambda_{max,1}$ and $\lambda_{max,2}$, ε_{21} and ε_{22} are the absorption coefficients of dye 2 at $\lambda_{max,1}$ and $\lambda_{max,2}$, respectively. *b* (cm) is the length of light path. C_1 and C_2 (mol L⁻¹) are concentrations of dye 1 and 2. The method described by Al-Degs et al. (2007) was used to estimate the absorption coefficient. The absorption coefficient for each dye was determined from the absorbance measurements of the dye at specific concentration. The slope of the linear regression about the concentration absorbance curves gave the value of the absorption coefficient for each dye at its wavelength of maximum absorption and at the wavelength of maximum absorption for the other dyes.

Results and discussions

Adsorption of dyes in the single and binary solution

Industry wastewater was normally a mixture of many compounds rather than a single one. The competitive adsorption of two dyes on fly ash was investigated using binary premature dye solutions. Figure 2 illustrated the mass of dye adsorbed on FA in single and binary solution. For the single solution, the adsorption of each dye was quite rapid in the first 50 min, and then a plateau formed after 50 min of contact. For the binary systems, the adsorption of disperse dye and anionic dye were a simultaneous process. However, the competitive adsorption showed clearly the interaction between dyes. It could be seen that the adsorption of all anionic dye (acid dyes and reactive dye) reached equilibrium within 50 min in binary solution, but the equilibrium time of disperse dye was longer than 50 min. It indicated that the adsorption rate of disperse dye (DR74) was affected apparently by the existence of the anionic dyes during the competitive adsorption while the rate of anionic dyes was almost unaffected by DR74 at all.

Table 2 presented the adsorption data of the single and binary systems. The adsorptive capacity of DR74 decreased from 3.899 mg g⁻¹ in single solution to 2.681, 3.378, 3.244, and 3.163 mg g⁻¹ in DR74/RB4, DR74/ RB171, DR74/AB1, and DR74/AB193 systems, respectively. Similar results could be seen for all anionic dyes. The reduction of equilibrium adsorption of individual dye in binary mixtures could be attributed to the competitive effects. However, the reduction extent varied in different binary systems. For disperse dye, the lowest adsorption capacity was found in DR74/RB4 solutions, and decreased about 31.3 % when compared to adsorption from the single solution. The highest competition between DR74 and RB4

Fig. 2 Adsorption dyes on fly ash in the single and binary solutions



Table 2 Equili

adsorption amor binary systems

brium unt of single and	Single systems	$q \ (\mathrm{mg \ g}^{-1})$	Binary systems	<i>q</i> (mg g	$q (\mathrm{mg g}^{-1})$		
				DR74	Other dye		
	DR74	3.899					
	RB4	3.601	DR74/RB4	2.681	3.093		
	RB171	3.011	DR74/RB171	3.378	1.185		

4.375

4.132

803

Other dye

13.9

39.7

10.4

11.6

Reduction extent (%)

DR74

31.3

11.4

16.8

19.6

could be the result of the similar molecular dimension data (Table 1) of the two dyes. The data also indicated a high degree of competition between DR74 and RB171 as the reduction in RB171 adsorption capacity was 39.7 % in the binary systems. An explanation could be given as disperse dye with no ionizable group was easily adsorbed in the large micropores near the surface. This resulted in the reduction in size of the larger pores, and then hindered the adsorption of RB171, which has a larger molecular volume. Acid dye and disperse dye have similarity affinity for fly ash, as the reduction in adsorption capacity was about 10-20 % in the binary systems.

AB1

AB193

It could also be inferred from Table 2 that the total mass of adsorbed dyes in binary systems was higher than that in single solution. For example, the maximum adsorption values of 3.899 and 4.375 mg g^{-1} for DR74 and AB1, respectively, in the single solution while the total adsorbed mass of the two dyes in the binary system was 7.164 mg g^{-1} (3.224 + 3.920, respectively). These results suggested that more active adsorption sites of FA were occupied in binary mixtures. The increment of total adsorption in binary solution was also reported in adsorption of dyes on activated carbon (Al-Degs et al. 2007; Walker and Weatherley 2000).

Desorption of dyes in the single and binary solution

In order to determine the predominant adsorption mechanisms, adsorbent samples were extracted after equilibrium. About 1.0 g of adsorbent was added to 100 mL of N,Ndimethylformamide and the mixtures were stirred at 150 rpm for 3 days. Samples were separated by centrifugation and dyes were recycled. The recovery efficiency of dyes was calculated and showed in Table 3. For disperse dye, the recovery efficiency exceeded 90 % in single solution and was found to be from 86.06 to 91.67 % in binary mixtures. As comparison, the values of four anionic dyes were in range 35.85-48.82 % in single solution and 4.05-8.05 % in binary systems. These results revealed the uptake of dyes on FA was the combined action of the chemisorption and physisorption. However, the predominant adsorption mechanisms were chemisorption for anionic dyes and physisorption for disperse dye (Noroozi et al. 2008). This chemisorption mechanism of anionic dye may be due to complexation between the negatively charged groups (D-SO³⁻) in anionic dyes and positively charged SiO_2 or Al_2O_3 groups on the surface of FA (pH_{pzc} = 8.4) (Bulut et al. 2008; Sun et al. 2010).

3.244

3.163

DR74/AB1

DR74/AB193

3.920

3.652

Adsorption kinetics

Adsorption kinetics show the dependence on the physical and chemical characteristics of the adsorbent material and adsorbate species. From Fig. 2, the adsorption process included two stages both in the single solution and in binary mixtures: the first rapid stage and the later slower stage. Using the experimental data, the adsorption rate of dye at different stage was calculated using the simple rate expression r = dC/dt by differential method. The initial adsorption rate at the two stages was given in Table 4. It could be observed that the adsorption rate of dye in rapid stage was about 5-15 times greater than those in the slower stage for the single solution, with one exception for RB171 in binary mixtures. For single solution, the maximum

Table 3	Recovery efficiency of
the dyes	from used adsorbents

Single system	Recovery efficiency (%)	Binary systems	Recovery efficiency (%)	
			DR74	Other dye
DR74	91.35			
RB4	35.85	DR74/RB4	88.15	4.05
RB171	42.17	DR74/RB171	91.67	5.81
AB1	43.35	DR74/AB1	86.06	6.77
AB193	48.82	DR74/AB193	88.75	8.05



Table 4 The initial adsorptionrate at the rapid and slow stagesin initial dye concentration of100 mg L^{-1}

Single system	$r_{\text{Rapid}} (\text{mg g}^{-1} \text{min}^{-1})$	$r_{\rm Slow} \ (\text{mg g}^{-1} \text{ min}^{-1})$	Binary systems	$r_{\text{Rapid}} (\text{mg g}^{-1} \text{min}^{-1})$	$r_{\text{Slow}} \ (\text{mg g}^{-1} \text{ min}^{-1})$
DR74	4.75	0.72			
RB4	5.84	0.93	DR74/RB4	3.50/5.62	0.58/0.85
RB171	7.67	0.45	DR74/RB171	3.96/5.95	0.31/0.16
AB1	7.56	0.83	DR74/AB1	2.88/5.80	0.26/0.65
AB193	5.45	0.65	DR74/AB193	2.84/4.85	0.29/0.68

adsorption rate in the rapid stage was observed to be 7.67 mg g⁻¹ min⁻¹ for RB171, while the minimum value was 4.75 mg g⁻¹ min⁻¹ for DB74. For binary mixtures, the adsorption rate of anionic dyes in rapid stage was higher than that of DR74, which resulted in the preferential adsorption of anionic dye. This phenomenon could also be seen in Fig. 2 that the mass of anionic dye adsorbed on FA was higher than that of disperse dye (DR74) at the first 20 times. In the slow stage, the adsorption rate of RB171 was the minimum one either in single solution or in binary mixtures, which may be explained by the fact that the largest volume of RB171 hindered the pore diffusion.

In order to understand the comprehensive adsorption mechanisms and interactions of the two dyes onto FA, the pseudo-first and second-order kinetic adsorption models were employed. The optimal modal was considered base on two indicators: correlation coefficient (r^2) and root mean square error (RMSE). The correlation coefficient was frequently used to decide whether the model represents correctly the experimental data. RMSE was defined as:

$$\text{RMSE} = \sqrt{\frac{\sum_{1}^{N} (q_{\text{exp}} - q_{\text{cal}})^2}{N}}$$
(3)

where q_{exp} and q_{cal} are the experimentally measured and modal prediction for the amount of dyes adsorbed, respectively. Smaller RMSE represents the equation can predict more accurately the experiment results.

The pseudo-first and second-order kinetic models are the most common model to study the adsorption kinetics of dyes on FA. The simplified linear form of first-order model was given by the Eq. (4) (Srivastava et al. 2006; Wasewar et al. 2008; Zhu et al. 2012).

$$\ln\frac{(q_{\rm e}-q)}{q_{\rm e}} = -k_1 t \tag{4}$$

where $q \pmod{g^{-1}}$ is the amount of material adsorbed at time t, $q_e \pmod{g^{-1}}$ is the adsorption capacity at equilibrium, $k_1 \pmod{q_e}$ (mg g⁻¹) represents the rate constant of the firstorder model. When $\operatorname{Ln}((q_e - q)/q_e)$ was plotted against t (plot not shown), Eq. 2 showed a straight line passing through the origin (Srivastava et al. 2006). Batch dynamic parameters were obtained using linear regression analysis (MATLAB 7.5) and showed in Table 5.

The second-order kinetic equation was defined as follows (Ho and McKay 1999; Srivastava et al. 2006):

$$q = \frac{k_2 q_{\rm e} t}{1 + k_2 t} \tag{5}$$

where k_2 (g mg⁻¹ min⁻¹) is the rate constant of the secondorder model. Batch dynamic parameters were obtained using linear regression analysis (MATLAB 7.5). Table 5 showed the adsorption parameters and the two indicators using each of the modal. The performances of kinetic models tested were comparable based on the correlation coefficients and RMSEs for each dye. For the five dyes in

Table 5 The first and second-order adsorption rate constants at initial dye concentration of 100 mg L^{-1}

	First-order kinetic model			Second-order kinetic model		
	$\overline{k_1 \ (\min^{-1})}$	r^2	RMSE	$k_2 (g mg^{-1} min^{-1})$	r^2	RMSE
DR74	0.038	0.968	0.478	0.114	0.989	0.165
RB4	0.033	0.929	0.641	0.151	0.991	0.103
RB171	0.028	0.957	0.570	0.085	0.984	0.151
AB1	0.054	0.922	0.506	0.205	0.993	0.109
AB193	0.052	0.949	0.419	0.181	0.981	0.226
DR74/RB4	0.028/0.041	0.983/0.962	0.130/0.392	0.036/0.146	0.985/0.997	0.105/0.057
DR74/RB171	0.032/0.049	0.963/0.955	0.321/0.367	0.065/0.369	0.987/0.997	0.127/0.024
DR74/AB1	0.034/0.041	0.967/0.953	0.151/0.479	0.052/0.137	0.997/0.989	0.117/0.062
DR74/AB193	0.038/0.063	0.961/0.971	0.298/0.378	0.070/0.148	0.978/0.981	0.164/0.261



single system, correlation coefficients for the first rate modal were in the range 0.922-0.968, while these values for the second rate modal were bigger than 0.98. At the same time, higher RMSE values (RMSE = 0.419-0.641) were obtained with the first-order kinetics. These results suggested pseudo second-order modal was applicable to the adsorption kinetics. Similar kinetic results were found for every dye in binary systems, even if there were competitive adsorptions between dyes. According to the higher r^2 and the lower RMSEs, the pseudo second-order showed better fitted to the kinetic data of dves on the FA in the mixtures solution, which indicated that the adsorption kinetic mechanism of dyes was unaffected by the competitive adsorption processes. However, the second-order kinetic rate constants of each dye decreased in binary system in contrast to those in single solution with only one exception for RB171. The values of the rate for RB171 increased $0.086 \text{ g mg}^{-1} \text{ min}^{-1}$ in single from solution to $0.369 \text{ g mg}^{-1} \text{ min}^{-1}$ in binary solution.

Adsorption rate controlling mechanism

Adsorption is a transient process involving some sort of Fick diffusion in both the fluid and adsorbent phases. It was important to identify the types of rate controlling mechanism. A number of mechanisms were suggested to control the adsorption rate of dye-adsorbent systems. Of the mechanisms mentioned above, external film and pore diffusion were most likely to prevail during dye adsorption onto fly ash.

According to previous studies (Kumar and Porkodi 2007; Wang and Li 2007), the intraparticle diffusion plot represented a multilinearity which characterized the two or more steps involved in the sorption process. The intraparticle diffusion model is defined as follows:

$$q = k_{\rm i} t^{0.5} \tag{6}$$

where $k_i (\text{mg g}^{-1} \text{min}^{-0.5})$ is the intraparticle rate constant. Figure 3 described the plot of q versus $t^{0.5}$ for dyes on FA. It was observed that there were two linear portions, indicating two adsorption stages. The initial sharper portion of the plot indicated the external surface adsorption or instantaneous adsorption stage while the second linear portion was due to intraparticle or pore diffusion. The kinetic parameters calculated from the diffusion kinetic models were shown in Table 6. As seen, a higher rate constant of intraparticle diffusion was for DR74 either in single solution or in binary solution. The values of DR74 increased in binary mixtures, although adsorption capacities at equilibrium decreased as compared to those in single system. However, the rate constants of the four anionic dyes were similar to those obtained in single solution. It was slightly lower for AB1 and a little higher for RB171.

In order to understand the adsorption rate controlling mechanism, a mathematical modal was used to gain the intraparticle diffusion coefficient. Intraparticle diffusion, $D \ (\text{cm}^2 \text{ s}^{-1})$, the sum of pore and surface diffusion, may simply be calculated from the following equation (Khraisheh et al. 2002).

$$-\log\left(1-\left(\frac{q}{q_e}\right)^2\right) = \frac{4\pi D}{2.3d^2}t\tag{7}$$

where d (µm) is the mean particle diameter. Table 6 showed the calculated results and revealed that the values of the internal diffusion coefficient fall well within the magnitudes reported in literature (Chazopoulos et al. 1993), specifically for chemisorption systems (10^{-5} to 10^{-13} cm² s⁻¹).

As illustrated in Fig. 3, the external mass of dye particles at initial time periods was significant for adsorption process. The external diffusion model (Findon et al. 1993; Rengaraj and Moon 2002) assumes that the concentration at the adsorbent surface tends to zero and the intraparticle diffusion is negligible at early times of contact. The change in concentration of adsorbates with respect to the time is related by the following equation (Kumar and Porkodi 2007):

$$-\frac{\mathrm{d}C}{\mathrm{d}t} = k_{\mathrm{f}}\frac{A}{V}(C_{\mathrm{t}} - C_{\mathrm{e}}) \tag{8}$$

where C_t , C_e , A/V, and t are the concentration (mg L⁻¹) at time t and at equilibrium, the total interfacial area of the particles (cm²) to the total solution volume (L) and adsorption time, respectively. A/V is expressed as Eq. (9) (Dizge et al. 2008).

$$\frac{A}{V} = \frac{3m}{\rho d} \tag{9}$$

where $m \text{ (g cm}^{-3})$ is the adsorbent dosage and ρ is the apparent density of the adsorbent (g cm⁻³). Using the initial conditions (i.e., $C_t = C_0$ and $C_e = 0$ at t = 0), therefore

$$\left[\frac{\mathrm{d}C/C_0}{\mathrm{d}t}\right]_{t=0} = -k_\mathrm{f}\frac{3m}{\rho d} \tag{10}$$

The external diffusion coefficient $k_{\rm f}$ can be obtained using differential calculation in MATLAB 7.5. The calculated external mass transfer coefficient, $k_{\rm f}$ (cm s⁻¹), for the adsorption of dye on FA was also given in Table 6. The external mass transfer coefficient of DR74 in single solution and binary mixtures was the minimum one. The value decreased from 1.65×10^{-4} cm s⁻¹ in single solution to $0.98 \times 10^{-4} - 1.29 \times 10^{-4}$ cm s⁻¹ in binary mixtures. The external mass transfer coefficients of the four anionic dyes are slightly lower in binary solution as compared to the single dye systems. By comparing the



Fig. 3 Intraparticle diffusion

kinetics for dyes on fly ash



Table 6The kinetic rateconstants of dyes on fly ash

	$k_{\rm i}$ (mg g ⁻¹ min ^{-0.5})	r^2	$D \times 10^{8}$ (cm ² s ⁻¹)	r^2	$k_{\rm f} \times 10^4$ (cm s ⁻¹)	B _N
DR74	0.034	0.978	0.33	0.968	1.65	217
RB4	0.027	0.952	0.28	0.942	2.05	319
RB171	0.022	0.995	0.23	0.965	2.42	263
AB1	0.026	0.997	0.30	0.930	2.36	150
AB193	0.021	0.995	0.39	0.957	1.79	195
DR74/RB4	0.053/0.027	0.981/0.920	0.22/0.32	0.976/0.987	0.98/2.01	188/265
DR74/RB171	0.094/0.026	0.942/0.929	0.22/0.28	0.935/0.976	1.39/2.24	475/292
DR74/AB1	0.038/0.024	0.952/0.923	0.28/0.34	0.963/0.959	1.09/2.02	357/251
DR74/AB193	0.069/0.021	0.981/0.922	0.28/0.46	0.943/0.989	1.02/1.62	156/149

values of k_i and k_f , it could be concluded that the competition adsorption in binary solutions unaffected the adsorption kinetic behavior of anionic dyes, but it improved the intraparticle rate and decreased the external diffusion coefficient of disperse dye in binary solution.

When the external and intraparticle diffusion coefficients are calculated for a given adsorption system, the Biot number can then be estimated from the Eq. (11) (Apiratikul et al. 2011; Queka and Al-Duri 2007):

$$B_{\rm N} = k_{\rm f} \frac{r}{D} = k_{\rm f} \frac{d}{2D} \tag{11}$$

The Biot number gives a criterion for the predominance of surface diffusion against external diffusion. As seen in Table 6, all Biot number obtained from all adsorption system was higher than 100, which suggested that



adsorption of dyes on FA was mainly controlled by intraparticle diffusion mechanism (Guibal et al. 1998; Khraisheh et al. 2002).

Conclusion

The competitive adsorption of disperse dye and anionic dye on fly ash was investigated. The following conclusions could be derived from the present study.

1. Adsorption equilibriums were reached within 50 min for all dyes used in the single solution. In binary solutions, the competitive adsorption unaffected the equilibrium time of all anionic dye (acid dyes and reactive dye), but increased the time of attaining equilibrium of disperse dye. Desorption of dyes suggested the predominant adsorption mechanisms were chemisorption for anionic dyes and physisorption for disperse dye.

- 2. Adsorption process included the first rapid stage and the later slower stage. For binary mixtures, the anionic dyes could be adsorbed preferentially on FA at the first stage. The initial adsorption rate of dye in rapid stage was about 5–15 times greater than those in the slower stage with one exception for RB171 in binary mixtures. The second-order kinetic models fitted better to the equilibrium data of all dyes either for the single solution or for the binary mixtures.
- 3. The maximum intraparticle diffusion rate constant and the minimum external mass transfer coefficient was found for disperse dye both in single and in binary solution. The intraparticle diffusion constants and external mass transfer coefficients of the four anionic dyes in the binary solution are similar to those obtained in single solution. The Biot number confirmed the intraparticle diffusion was the rate-limiting step in dye sorption process.

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