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Removal of hexavalent chromium from aqueous solution by crosslinked mangosteen peel biosorbent

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Abstract In present study the potential of a new biosorbent obtained from mangosteen peel was investigated for the removal of hexavalent chromium and trivalent chromium from the aqueous solutions. The biosorbent was modified by using concentrated sulphuric acid as the crosslinking reagent. Various batch experiments were carried out testing different values of several parameters such as pH, contact time, adsorbent dosages, and metal ion concentration to identify the optimum conditions. The effect of these parameters on the adsorption capacity and selectivity was studied. Results showed that optimum pH for hexavalent chromium adsorption was 3.0 and adsorption equilibrium was attained in 120 min. The adsorption data were found to fit Langmuir model and pseudo-secondorder equation. The adsorption mechanism for hexavalent chromium was found to be a complex process involving esterification, reduction and ion exchange, while cation exchange is the main route for trivalent chromium adsorption. From the above results, the crosslinked mangosteen peel gel appears to be a potential low-cost and effective adsorbent for hexavalent chromium and trivalent chromium removal from the waste water.

Keywords Biosorption · Chromium · Mangosteen peel · Concentrated sulphuric acid

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Introduction

Chromium, along with lead, cadmium, arsenic and mercury, is usually ranked as the first-category poisonous chemicals that should be strictly controlled during their usage and discharging (Levankumar et al. 2009; Kratochvil and Volesky 1998; Memon et al. 2009). Chromium usually exists in aqueous medium in +III and +VI oxidation states; as to Cr(VI), it is much more toxic than the trivalent chromium species, and may cause serious damage to kidney, liver and stomach (Inoue et al. 2010; Sarin and Pant 2006). More serious is that Cr(VI) has been proved to be associated with lung cancer (Mohan et al. 2005). The main source of chromium in the environment is largely from the uncontrolled release of chromium-related industries, such as leather tanning, electroplating, chromium smelting and so on (Gupta et al. 2010). So it is very important to regulate and restrict the whole process of production, usage, and discharging of chromium. The maximum permitted concentration of hexavalent chromium in drinking water suggested by United States Environmental Protection Agency (US EPA) and World Health Organization (WHO) is less than 0.05 mg L^{-1} (USEPA 2011; WHO 2004).

The most common Cr(VI) removal technology applied in waste water is reduction by FeSO₄ coupled with coagulation. This method has prominent disadvantages such as production of toxic sludge requiring tedious treatment, excessive amounts of FeSO₄ and alkaline reagents left in the treated water (Gupta et al. 2010; Jain et al. 2010; Memon et al. 2009; Chand et al. 2009a, b). The other techniques such as reverse osmosis and electrolysis as well as ultrafiltration (Volesky 2007; Demirbas 2008), ion exchange (Gupta et al. 2010; Memon et al. 2009) and adsorption on activated carbon (Chand et al. 2009b; Mohan et al. 2005), may have the problems in high operation and



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maintenance cost, or poor removal efficiency. So it is necessary to develop more cost-effective techniques to remove Cr(VI) from the waste water.

In recent years, biosorption has attracted many people's interest in water treatment (Ali and Gupta 2006; Ali 2010, 2012; Dhakal et al. 2005; Ghimire et al. 2008a, b; Inoue et al. 2010; Parajuli et al. 2009; Volesky 2007); biosorbents are the key material determining the purification efficiency, operation cost and practical feasibility. In the early periods, fungi and algae were the most common raw materials considered for biosorption, and adsorption mechanisms of various elements were demonstrated in these early studies (Pavasant et al. 2006; Volesky 2007). However, more abundant and easier available sources of biomass for biosorbents preparation, in many countries, are the biowastes from the agro-industry. Huge amounts of agro-wastes provide sufficient raw materials available as much cheaper and more efficient biosorbents. Many kinds of biowastes, as long as the people can think of it, such as orange juice residue (Dhakal et al. 2005), apple juicing residue (Ghimire et al. 2008a), persimmon peel (Inoue et al. 2010), have been tested for their possibility to be used as the biosorbents for removing toxic elements from drinking water, exhibiting exciting prospective for application as biosorbents. For Cr(VI) removal from drinking water, many biosorbents have been evaluated about their adsorption behavior, and most of them were found to give rise to adsorption coupled with the reduction mechanism, or direct adsorption by esterification process (Chand et al. 2009a, b; Jain et al. 2010; Levankumar et al. 2009; Mohan et al. 2005). Those biomasses containing polyphenols typically demonstrated above-mentioned adsorption behavior (Chand et al. 2009a, b; Inoue et al. 2010), so this gave us an idea to search for the abundant biomaterials to prepare the sorbent, which will have excellent adsorption affinity specially to Cr(VI). Based on the above idea, in an earlier study we have prepared a kind of sorbent by using mangosteen peel as the raw material (Huang et al. 2013), which is rich of polyphenolic compounds such as anthocyanin, resveratrol etc. (Naczk et al. 2011; Tao et al. 2009; Wang et al. 2012; Zadernowski et al. 2009; Zein et al. 2010), through modification by alkaline soaking, i.e., saponification process. The obtained saponified mangosteen peel gel showed quite excellent adsorption properties for Cr(VI) removal with high selectivity and large adsorption capacity. The detailed description has been published (Huang et al. 2013), and in present work, we wanted to try another new modification method different from the previous one, which may help to remove low-molecular-weight and water-soluble components in the mangosteen peel more effectively by dehydration with concentrated sulfuric acid, and provide another modification method to make good use of mangosteen peel waste as biosorbent for water treatment.



This study was conducted from September 2011 to July 2013 in the State Key Laboratory of Advanced Metallurgy and school of metallurgical and ecological engineering, University of Science and Technology Beijing, Beijing, China.

Materials and methods

Preparation of the adsorbent

The mangosteen fruit used in this work was purchased in the local market of Jianxianggiao fair of Beijing city; it was imported from Thailand, and the peel used for study was obtained after its edible part was consumed. Typical preparation procedure of the adsorption gel was as follows. Fifty grams of wet mangosteen peel were crushed and the obtained purple pulp was dried in an oven at 60 °C for 12 h. Then the dried powder was mixed homogeneously with 50 cm³ concentrated sulphuric acid (96 %) and kept it in the atmosphere for 24 h in a glass beaker to allow condensation crosslinking reaction between the various polyphenol and polysaccharide hydroxyl groups in the mangosteen peel. The obtained black residual solid was neutralized by sodium hydroxide solution and then filtered and reconditioned with 1 mol L^{-1} H₂SO₄ solution. After stirring for 30 min, the suspension was filtered and washed by hot distilled water several times until the pH of the filtered liquor approached 7.0. The wet filtered cake was dried in an air convection oven at 60 °C for 12 h, and the particle size was sieved to 100-200 µm. The final product of crosslinked mangosteen peel, designated as CMP, is used for the following adsorption tests.

Analysis of Cr(VI) and total chromium in the test solutions

For the determination of Cr(VI) concentration in the aqueous solution, the standard diphenylcarbazide method is used (Zhang and Chen 2012). The concentration of Cr(VI) before and after adsorption were analyzed with a ultraviolet–visible spectrophotometer (UNICO-2100, USA) at 540 nm wavelength by using the pink complex with diphenylcarbazide in acidic solution. The total chromium concentration in the filtrated liquor was measured by a Shimadzu model AAS-6800 atomic absorption spectrophotometer. So based on the above two analysis methods, Cr(III) concentration can be obtained easily from the difference between the total chromium and the Cr(VI) concentrations.

Batch wise adsorption tests

In batch wise adsorption tests, 25 mg of CMP dry gel and 15 cm³ of the individual metal solution at 1.0 mmol dm⁻³, were mixed and shaken at 293 K for around 24 h to attain equilibrium in a thermostatic air bath incubator reciprocating at 140 rpm. The initial pH values of the metal ion solutions were adjusted with 0.1 mol dm⁻³ HCl and 0.1 mol dm⁻³ NaOH solutions, and the pH values during the adsorption process were kept almost constant by using 0.1 mol dm⁻³ HEPES (2-[4-(2-hydroxylethyl)-1-piperazinyl] ethanesulfonic acid) as the buffer solution for each test sample. After adsorption, the metal ion solution samples were filtered and the metal concentration in the filtrated liquor was measured for evaluate the adsorption efficiency. The amount of adsorption for metal ions was calculated according to the following equations.

$$A = (C_i - C_e)/C_i \times 100 \tag{1}$$

$$Q = (C_i - C_e) \times V/W \tag{2}$$

where A is the adsorption percentage of metal ions on the gel (%), Q denotes the amount of metal ion adsorbed per unit mass of the adsorbent (mol kg⁻¹ or mol kg⁻¹), V is the solution volume (dm³), W is the dry mass of the adsorbent (kg or g), C_i and C_e represent the metal ion concentrations in the solution before and after adsorption (mol dm⁻³ or mg dm⁻³), respectively.

Reusability test of the eluted adsorbents

To check the reusability of the adsorbent, CMP adsorbent was subjected to repeated adsorption and elution experiments as follows. Twenty five milligrams CMP gel were shaken with 15 cm⁻³ 1 mmol dm⁻³ CrO₄²⁻ solution at pH 1 for 24 h, then the mixture was filtrated and the residue gel was washed by double-distilled water three times to remove the physically adsorbed ions. The metal ion-loaded gel was then shaken with 25 cm⁻³ 0.5 mol dm⁻³ hydro-chloric acid in a conical flask at 293 K for 2 h. Above adsorption and elution experiments were repeated, and the metal ion concentration before and after adsorption was measured each time to evaluate the adsorption efficiency.

Results and discussion

Characterization of the adsorbent

Figure 1 shows the SEM photo of the surface of crosslinked mangosteen peel gel. It is clear that the surface morphology of the gel is rough which is favorable for metal ions to be trapped and adsorbed on the gel. It also looks not



Fig. 1 Scanning electron micrograph of the surface of CMP gel

so compact, which will be helpful for metal ions to diffuse or penetrate into the inner parts of the adsorbent particles and good for the fast adsorption.

Figure 2 shows the FTIR spectra of CMP gel before and after adsorption of Cr(VI) at pH 3.0. The broad absorption peak at 3.345 cm^{-1} can be attributed to the O–H stretching vibration due to inter and intra-molecular hydrogen bonding of alcohols, phenols and carboxylic acids, and the absorption peaks at 2,970 and 2,932 cm^{-1} can be assigned to the C-H stretching of non ionic carboxylic groups, while peaks at $1,700 \text{ cm}^{-1}$ corresponds to the asymmetric and symmetric stretching vibration of C=O bond in ionic carboxylic groups of -COO-, respectively. The peaks at $1,450 \text{ cm}^{-1}$ (C=C bond), $1,286 \text{ cm}^{-1}$ (O–H bond) and 883 and 619 cm^{-1} (C–H bond) can be assigned to the phenolic group (Inoue et al. 2010), and they are found to become weaker or even disappear after adsorption of Cr(VI), indicating the effective adsorption of chromium ions on the CMP gel.

Effect of pH

The adsorption behavior of the saponified mangosteen peel gel for hexavalent and trivalent chromium at different equilibrium pH values is shown in Fig. 3. It can be found that two species of chromium demonstrate quite different adsorption patterns. It is clear that the CMP gel exhibits high selectivity for Cr(VI) over Cr(III) at pH <4, while at pH >4, the high selectivity was conversely changed for Cr(III) over Cr(VI). Obviously, the solution pH plays a very key role in this adsorption process. Quite similar adsorption behavior has also been reported for the polyphenol-containing biosorbents such as persimmon peel gel, grape peel gel and saponified mangosteen peel gel (Inoue et al. 2010; Chand et al. 2009a; Huang et al. 2013). The contrasting adsorption behavior for Cr(VI) and Cr(III) can



Fig. 2 Fourier transform infrared (FTIR) spectra of crosslined mangosteen peel (CMP) gel before and after adsorption of chromium ions at pH 3.0





Fig. 3 Effect of equilibrium pH on the percentage adsorption of Cr(VI) and Cr(III) ions on CMP gel. (Dry weight of the gel = 25 mg, volume of solution = 15 cm^3 , initial concentration of metal ions = 1 mmol dm⁻³, contact time = 4 h, temp. = 293 K)

be ascribed to the different uptake mechanisms during their adsorption process. Usually, according to the related research work (Inoue et al. 2010; Chand et al. 2009a), it can be proposed that adsorption of Cr(VI) may occur through a direct esterification reaction between the $HCrO_4^{-}$ and the phenol groups on the gel, and then the captured Cr(VI) might be reduced to Cr(III) by the electron rich polyphenolic ring in the gel and the produced Cr³⁺ would be adsorbed onto the CMP gel at high pH and desorbed at low pH (Inoue et al. 2010; Chand et al. 2009a; Huang et al. 2013), as shown in Fig. 3. So we can see that adsorption of Cr(VI) may include the following three steps, i.e., esterification, reduction, and uptake of produced Cr^{3+} . As described in many studies (Inoue et al. 2010; Chand et al. 2009a, b; Huang et al. 2013), H⁺ is favorable for the esterification and reduction steps while impedes the uptake of Cr^{3+} for it is believed to proceed according to a cation exchange mechanism. At low pH less than 3, as seen in Fig. 3, Cr^{3+} is hardly adsorbed by the CMP gel, so it can be concluded that the contribution to the removal of Cr(VI) in this pH range is mainly from the esterification process. And at lower pH, more H⁺ will promote more Cr(VI) to be transformed into Cr(III), so the A % (percent adsorption) will decrease in this acidic solution, and at pH higher than 3, the esterification process will be inhibited due to the high concentration of OH⁻, leading to the rapid decrease of A % at pH higher than 3.0 as shown in Fig. 3. So actually there is a competitive interaction between the esterification and reduction, and H⁺ plays a key role in the adsorption process of Cr(VI).

In order to elucidate the reduction mechanism of Cr(VI) during the adsorption onto the CMP gel, both Cr(VI) and Cr(III) concentrations in solution were measured as shown in Fig. 4. In this experiment, an initial solution containing only Cr(VI) was used to contact with CMP gel, and sampled for measurement after a certain time interval. The total chromium concentration in solution was measured by an atomic absorption spectrophotometer, and the Cr(VI) concentration was measured by the diphenylcarbazide method. The difference of total chromium and Cr(VI) was the Cr(III) concentration. Figure 4 shows that Cr(VI) concentration will decrease rapidly when adsorption occurs, while the concentration of Cr(III) will increase quickly from zero to a constant. In this experiment, there's no trivalent chromium in the initial solution, but it appears in the adsorption process, which proves the occurrence of reduction. This figure also shows that the final chromium concentration was actually the concentration of Cr(III) which has negligible adsorption onto the CMP gel at pH 3.0. Based on above analysis, it can be seen that the adsorption mechanism of Cr(VI) onto CMP gel is quite complex and different from that of the Cr(III) adsorption process.



Fig. 4 Time variation of the concentrations of Cr(VI) and Cr(III) as well as total chromium during adsorption. (Dry weight of the gel = 25 mg, volume of solution = 15 cm^3 , initial concentration of metal ions = 1 mmol dm⁻³, pH = 3.0, temp. = 293 K)



Fig. 5 Effect of temperature on the percentage adsorption of Cr(VI) on CMP gel. (Dry weight of the gel = 25 mg, volume of solution = 15 cm³, initial concentration of metal ions = 1 mmol dm⁻³, pH = 3.0, temp. = 293 K)

Effect of temperature

Figure 5 shows the effect of temperature on the % adsorption of Cr(VI) onto CMP gel at pH 3. It can be found that the adsorption capacity will decrease when temperature increases, which suggests that the adsorption of Cr(VI) is an exothermic process. It was further found that above adsorption data can be described by pseudo-second-order kinetics equation, as shown in Fig. 6. Based on the results, an Arrhenius formula was used to evaluate the apparent activation energy to be 46.93 kJ mol⁻¹ by plotting



Fig. 6 Pseudo-second-order kinetics plots at different temperatures for Cr(VI) adsorption on CMP gel (conditions as Fig. 5)

In*K* versus 1/*T*, as indicated in Table 1; this suggests that the key step to limit the whole adsorption process is a chemical reaction, probably esterification and reduction steps. In fact the scale value of activation energy gives an indication on the mechanism of the adsorption process: values higher than 12 kJ mol⁻¹ indicate that chemical reaction is the controlling step, whereas values lower than 12 kJ mol⁻¹ suggest that diffusion is the controlling step (Li 1992). The adsorption activation energy value is quite close to that of the saponified mangosteen peel gel with 42.73 kJ mol⁻¹ (Huang et al. 2013). Then $\ln K_c$ was plotted versus 1/T (Table 1), and from the fitting line the enthalpy of the adsorption process was evaluated to be -86.95 kJ mol⁻¹, indicating that the adsorption of Cr(VI) onto CMP gel is an exothermic process.

Effect of gel dosage

Figure 7 shows the effect of CMP gel dosage on the % adsorption of Cr(VI). It can be seen that only 1 g dm⁻³ can make the % adsorption to reach a constant. Though the initial concentration of Cr(VI) is only 2.7 mg dm⁻³, the residual chromium concentration is still kept at 0.5 mg dm⁻³ even adding more gel in the solution. It is clear that at pH near 3.0, most of the Cr(VI) will be uptaken by the esterification process, while a part of Cr(VI) will also be reduced to Cr(III) which is hardly adsorbed by CMP gel in this case and is left in the residual solution as trivalent chromium. The reduced amount of Cr(VI) into Cr(III) is mainly determined by solution pH when the CMP gel is sufficient, so the residual chromium concentration had the same values whatever the gel dosage.



Table 1 Calculation (alloss of apparent activation energy and enalary) of the adoctpion process						
Equation models	Fitting model of the data	Calculation energy values, kJ mol ⁻¹	Correlation coefficient, R^2	Remarks		
$\ln K$ vs $1/T$	y = -5,645x + 12.11	46.93	0.909	Apparent activation energy for adsorption		
$\ln K_c$ vs $1/T$	y = 10,458x - 34.29	-86.95	0.939	Enthalpy value of adsorption		

Table 1 Calculation values of apparent activation energy and enthalpy of the adsorption process



Fig. 7 Effect of dosage on the percentage adsorption of Cr(VI) on CMP gel. (Dry weight of the gel = 25 mg, volume of solution = 15 cm³, initial concentration of metal ions = 2.7 mg dm^{-3} , pH = 3.0, temp. = 293 K, contact time = 4 h)

Adsorption isotherms for Cr(VI)

The isotherm test was made to evaluate the adsorption capacity of the CMP gel to uptake Cr(VI) from aqueous solution. Isotherm tests were carried out batch wisely using various concentrations of Cr(VI) ranging from 0.1 to 10 mmol/L at solution pH 3. The adsorption isotherms on the saponified mangosteen peel gel are shown in Fig. 8a, from which the adsorption of Cr(VI) seems to occur according to a Langmuir-type adsorption, i.e., the adsorbed amount increases with increasing concentration of metal ion at low metal ion concentration while it tends to reach a constant value at high metal ion concentration.

So the adsorption isotherms were evaluated using Langmuir model (Inoue et al. 2010) represented as Eq. (3), as depicted in Fig. 8b.

$$C_e/Q_e = 1/(bQ_m) + C_e/Q_m \tag{3}$$

where $Q_e \pmod{\text{kg}^{-1}}$ or $g \text{kg}^{-1}$ is the mass of adsorbed metal ions per gram of adsorbent, C_e (mol L⁻¹, or denoting as M) is the concentration of metal ion in aqueous solution at equilibrium, Q_m (mol kg⁻¹) is the maximum capacity of adsorbent, and b (L kg⁻¹) is the binding constant.



Table 2 lists the maximum adsorption capacities of the various adsorbents including CMP gel. This table shows that the



 Table 2 Comparison of the maximum adsorption capacity for Cr(VI) on different adsorbents

Adsorbents	pН	Temp. (°C)	Adsorption capacity (mol kg ⁻¹)	References
Persimmon tannin	2.0	30	5.27	Nakajima and Baba (2004)
Ocimum americanum L. seed pods	1.5	27	1.60	Levankumar et al. (2009)
Crosslinked persimmon gel	1.0	30	7.18	Inoue et al. (2010)
Crosslinked grape gel	4.0	30	1.91	Chand et al. (2009a)
Barley straw	2.0	30	1.68	Chand et al. (2009b)
Saponified mangosteen peel	3.0	20	1.99	Huang et al. (2013)
Crosslinked mangosteen peel	3.0	20	2.00	This study

polyphenol-bearing biosorbents have higher adsorption capacities than other biosorbents such as seed pods and barley straw. It is also found that at pH 3, the mangosteen peel saponified with alkaline reagent (Huang et al. 2013), and crosslinked by concentrated sulfuric acid (present study) have quite close adsorption capacity for Cr(VI), which suggests that these two modification methods both effectively activated the adsorption ligands contained in the mangosteen peel. As compared to other adsorbents, the preparation of CMP gel is quite simple and economic, just using mangosteen peel and a small amount of concentrated sulfuric acid as the crosslinking reagent, which indicates great potential for its application to cost-effective removal of Cr(VI) from various industrial effluents.

Reusability of the eluted adsorbents

To check the reusability of the CMP gel, CMP adsorbent was subjected to repeated adsorption and elution experiments. The adsorption capacity of the CMP gel was found to be practically kept more than 90 % after 10 cycles of use, so it can be seen that multiple use of the adsorbent is feasible.

Conclusion

An excellent adsorbent for Cr(VI) removal was prepared by using mangosteen peel as the raw material through a crosslinking process with concentrated sulphuric acid as the modifier. The synthesis process is simple, economical and non-poisonous, and the prepared adsorbent shows high affinity, large loading capacity for Cr(VI), which might be an alternative to the costly resins or activated carbon for the removal of Cr(VI). The adsorption isotherm indicates that the removal of Cr(VI) can be described by a Langmuir model, and the maximum adsorption capacity for Cr(VI) at pH 3.0 was evaluated to be 2.0 mol kg^{-1} for Cr(VI). The adsorption of Cr(VI) on CMP gel may take place according to the following two routes, i.e., one is through an esterification reaction, and the other is a reduction of the esterified Cr(VI) into Cr(III) by the polyphenol components contained in the mangosteen peel followed by an adsorption of reduced Cr(III) onto the gel. So an optimal pH actually depends on competitive result of above two adsorption processes which are both highly pH-dependent. For its low cost and large adsorption capacity for Cr(VI) under acidic conditions, CMP gel prepared by the proposed process might find good application in the treatment of various Cr(VI)-bearing waste water.

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