ORIGINAL PAPER



# Investigation of anaerobic biodegradability of real compost leachate emphasis on biogas harvesting

H. Hashemi<sup>1</sup> · A. Ebrahimi<sup>2</sup> · A. Khodabakhshi<sup>3</sup>

Received: 9 September 2014/Revised: 14 March 2015/Accepted: 21 April 2015/Published online: 12 May 2015 © Islamic Azad University (IAU) 2015

**Abstract** Anaerobic treatability of real compost leachate was assessed using laboratory-scale anaerobic sequencing batch reactor at mesophilic conditions. Interventional study was conducted at wide range of organic loading rate 0.93–25 g  $l^{-1}$  day<sup>-1</sup> by varying hydraulic retention times 23 and 12 h. Initial chemical oxygen demand (COD) was 1.85-25 g l<sup>-1</sup>. pH variations; total, soluble (SCOD), readily biodegradable (rbCOD) chemical oxygen demand; volatile fatty acids degradation; biogas production; and methane fraction were considered in this study. The organic matter removal efficiencies were in the range of 76-81 % depending on loading rates applied. The maximum volumetric methane production rate of 5.7 l CH<sub>4</sub> - $1^{-1}$  day<sup>-1</sup> was achieved at the loading rate of 19.65 g l<sup>-1</sup> day<sup>-1</sup>. About 85 % of removed organic matters during the biodegradation were converted to the methane. The results have shown that the anaerobic batch reactor could be an appealing option for changing compost leachate into the useable products such as biogas and other energy-rich compounds, which may play a serious role in meeting the world's ever-increasing energy requirements in the future.

A. Khodabakhshi khodabakhshi16@gmail.com

- <sup>2</sup> Environmental Sciences and Technology Research Center, Department of Environmental Health Engineering, Shahid Sadoughi University of Medical Sciences, Yazd, Iran
- <sup>3</sup> Environmental Health Engineering, School of Health, Shahrekord University of Medical Sciences, Shahrekord, Iran

**Keywords** Batch process · Compost liquid · Methane · Organic load

# Introduction

Improper source separation of waste in developing countries like Iran leads to the production of considerable waste with high water content and various pollutants (Troschinetz and Mihelcic. 2009). Putrescible wastes as a raw material of compost consist of high water content, and water is also produced during biodegradation processes. This wastewater seeps out and forms what is known as leachate (Khalid et al. 2011). For example, Isfahan compost plant produces leachate at an average of about 40 m<sup>3</sup> day<sup>-1</sup>. Leachates are heavily polluted wastewaters with a complex composition containing microorganisms typically Bacillus, Citrobacter, Enterobacter, E. coli, Klebsiella, Nieseria, Pseudomonas, Shigella, Staphylococcus, Streptococcus and Vibrio (Obire and Aguda. 2002); dissolved organic matters; inorganic macrocomponents; heavy metals; and xenobiotic organic compounds with a foul odor (Safari et al. 2011). It leached from several points of compost facilities and viewed as a pollution source of soil, surface water and groundwater, on account of organic substrates and potential pathogens which pose a risk to the environment and public health if released without proper treatment (El-Fadel et al. 2003; Zhang et al. 2007). In spite of compost production, the management and effective treatment of compost leachate is worth paying attention. Compost leachate may be treated anaerobically, saving environment and converting the organic material partially to biogas (Bonmat et al. 2001). Even toxic compounds may be degraded anaerobically depending on the process applied. One prerequisite is that a feed waste contains a



<sup>&</sup>lt;sup>1</sup> Environmental Health Engineering, School of Health, Shiraz University of Medical Sciences, Shiraz, Iran

considerable amount of organic matter that is finally converted mainly to CH<sub>4</sub> and CO<sub>2</sub>. At least 20 % of energy used in the EU comes from renewable sources, and 10 % of the fuels used in transport being biofuels (Karim et al. 2005). In this context, anaerobic digestion with biogas generation is a good answer to today's environmental challenges (Ghani and Idris. 2009). Indeed, anaerobic digestion with energy recovery produces less greenhouse gases than incineration or landfilling (Khalid et al. 2011). As a rule of thumb, wastes containing <60 % of volatile solids are rarely considered as substrates for anaerobic digestion. Therefore, with regard to the total solids content, the percentage of volatile solids, the C:N ratio and the biodegradability, compost leachate can be used as suitable substrate for anaerobic digestion that generates useable biogas as a source of energy (Agdag and Sponza. 2005). Few studies have been conducted on the anaerobic digestion of compost leachate. Furthermore, most of those studies use synthetic leachate. Compost leachate and young landfill leachate normally contain high amounts of volatile fatty acids (VFAs). These readily biodegradable matters account for the bulk of the chemical oxygen demand (COD) of leachate, so the ratio of biological oxygen demand (BOD) to COD is relatively high (Nayono et al. 2010). The anaerobic treatment of young leachate in an anaerobic system allows the anaerobic stabilization to terminate which was initiated in the tip. Among the biological processes found in the scientific literature, ASBR have shown good removal efficiencies in pretreatment of municipal solid waste compost plant leachate (Amin et al. 2014). COD removal was as high as 85–90 % for the whole sequential anaerobic-aerobic treatment process, while the COD removal efficiency in the anaerobic stage was only 60 % (Palatsi et al. 2011). HRT of about 70 days would be needed to achieve 90 % COD removal in a leachate containing in excess of about 150,000 mg  $1^{-1}$  of COD for which relatively large reactors are required (Oz and Yarimtepe. 2014). Nevertheless, it has been claimed that the anaerobic digestion of leachate does not remove ammoniacal N at all (often up to 1000 mg  $l^{-1}$ ) and indeed is more likely to increase concentrations of this main contaminant of leachates. Sometimes, co-digestion of leachate and wastewater can be an attractive option for upgrading digestion efficiency, providing balance of nutrients and the synergy effect between organic substrates (Xiaofeng et al. 2014). According to the Bouallagui et al. (2009) research, during the digestion of fruit and vegetable waste, additions of abattoir wastewater or activated sludge could enhance biogas yields by 51.5 and 43.8 %, respectively. The added wastewater or sludge lowered carbon-nitrogen ratio and enhanced biogas yields (El-Mashad and Zhang 2010). The biogas produced through the anaerobic digestion of organic fraction of municipal solid waste comprised of 50-70 vol % CH<sub>4</sub>, 30–50 vol % CO<sub>2</sub> and trace volatile compounds that should be controlled and reduced to ultra-low levels in a gas-cleaning section (Mata-Alvarez et al. 2000; Khalid et al. 2011). This study will mainly focus on the real compost leachate COD removal and biogas production using ASBR process in Isfahan (Iran) during the period of June 2012 and March 2014.

# Materials and methods

## **Pilot setup**

The total volume of reactor was 2 l with 0.5 l filled with sludge, 1 l was completed with the addition of pretreated leachate, and 0.5 l was used as freeboard for biogas storage into the headspace. The laboratory-scale batch reactor consisted of a thermostated water bath glass container with a liquid working volume of 20 l which was maintained a temperature of 37 °C. To obtain a homogeneous suspension, feeding was injected by one etatron pump through the bottom of the reactor. The reactor content was mixed with a vertical mixer inside the reactor. At the end of each cycle, mixer was stopped and after 1-h settling, and 1 l of supernatant was withdrawn from decant valve. A detail of the ASBR reactor is presented in Fig. 1.



**Fig. 1** Simplified schematic diagram of the ASBR pilot: *1* feeding tank, *2* injection pump, *3* sludge intake valve, *4* sampling valve, *5* decant valve, *6* ASBR, *7* mixer, *8* bath container water, *9* Tedlar bag, *10* gas meter, *11* PLC

#### ASBR start-up and operational pattern

Experimental study was conducted at wide range of organic loading rate (OLR = 0.93-25 g COD  $l^{-1}$  day<sup>-1</sup>) by varying hydraulic retention times (HRT = 23 and 12 h) and initial COD of 1.85–25 g  $1^{-1}$ . Because of the seasonal biodegradability variations and acidic nature of the compost leachate, by using higher HRT and dilution of the influent, anaerobic reactor was adopted gradually. So there was no need for supplementation alkalinity. The reactor was operated with a 24-h cycle approximately consisting of fill (10 min of mixed fill), react (23-12 h), settle (1 h), and decant (10 min) phases. Partially granulated anaerobic sludge with a total VSS amount of 7500 mg  $l^{-1}$  was taken from the methanogenic digester of wastewater treatment plant in Isfahan (Iran) and used as seed in ASBR reactor. Sometimes, biomass loss was occurred due to influent shock that was replaced, using 0.5 l acclimated sludge. In first runs, HRT and OLR were fixed. The influent rate was kept constant at 1 l day<sup>-1</sup>. After steady state, HRT was decreased and OLR was increased. Adaptation of microorganisms to the leachate and operating conditions was also pointed out along the consecutive batches.

#### Experimental setup and analytical methods

ASBR reactor was inoculated with 1 l sludge from an anaerobic digester that operated at 35 °C. Washing of sludge with distilled water to remove coarse particles and then sieving with a pore diameter of 5 mm were performed. The sludge characteristics are given in Table 1.

Operational issues of the each cycle including pH, total COD, rbCOD, SCOD (in filtered samples) and TSS were analyzed in accordance with standard methods twice a week. The pH was measured using a calibrated pH meter (Schott, Model GC 824). COD measurement was conducted based on Dichromate method (closed reflux, 5220C, colorimetric method) (APHA 2005). The floc/filtration method was used for measurement of the rbCOD concentration (Agdag and Sponza. 2005). The reactor was properly sealed to ensure anaerobic condition and to prevent gas leakage. After sealing the gas-tight lids on the reactors, the produced biogas in the ASBR was collected using 50-ml plastic bag, and the daily volume was read directly using the ELSTER wet-test gas meter, PVC model (Elster,

Table 1Sludge characteristicwas used to ASBR reactor

Parameter	Value	Unit
TSS	10,700	mg/l
VSS	7950	mg/l
VSS/TSS	0.74	-
pH	7.55	-

AMCO, Germany). Composition of the produced biogas was determined by gas chromatograph (Auto System Perkin Elmer, USA) that was equipped with a packed column (Perkin Elmer,  $6' \times 1.8''$  OD, 80/100, Mesh, USA) and a thermal conductivity detector (Perkin Elmer, USA). An inject temperature of 150 °C was applied, and the carrier gas was nitrogen operated with a flow rate of 20 ml min<sup>-1</sup> at 75 °C. An Agilent technologies system consisting of a 5975 °C Inert MSD with a triple axis detector equipped with a 7890A gas chromatograph with a split/splitless injector was used for the quantification and confirmation of the VFA.

#### **Results and discussion**

#### Feeding leachate characterization

The mixture of leachate produced in different units of Isfahan compost plant (Receiving hall, Shredding, Press and Fermentation site) was obtained. Compost leachate was pretreated by anaerobic mixing baffled reactor (AMBR) to alleviate organic loading and solids, and after that, it was injected to the ASBR. Characteristics of the feed are presented in Table 2.

The leachate showed high COD and BOD<sub>5</sub> concentrations. The BOD<sub>5</sub>/COD ratio (or biodegradability factor) varied between 0.28 and 0.38, and all samples had a relatively high biodegradability index. Total suspended solids concentration was high due to the leachate being fed from the anaerobic reactor. The pH of the leachate samples was varying between 6.4 and 7.5. The leachate had a strong odor, probably due to a high content of volatile fatty acids. Also compost leachate contains different microbial communities.

## **COD** removal

Figure 2 shows the removal efficiencies of COD, SCOD and rbCOD parameters in ASBR reactor at various OLRs. Influent COD, SCOD and rbCOD concentrations were in

Table 2 H	Results of	of the	main	feed	leachate	parameters	analyzed
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Parameter	Range	Mean $\pm$ SD	
Total COD (g/l)	1.85-25	$11.19 \pm 8.64$	
SCOD (g/l)	0.42-7.25	$3.12 \pm 2.45$	
rbCOD (g/l)	0.07-2.3	$0.88\pm0.86$	
BOD <sub>5</sub> /COD	0.28-0.38	$0.33 \pm 0.03$	
Ph	6.23-8	$7.38\pm0.42$	
EC (mS/cm)	2.13-20.68	$11.57 \pm 7.74$	
TDS (g/l)	1.49–11.02	$6.85\pm4.05$	





Fig. 2 Removal efficiency of COD, SCOD and rbCOD in ASBR reactor at different OLRs during the periods of operation

the ranges of 1.85–25, 0.42–6.52 and 0.07–2.3 g  $l^{-1}$ , respectively. In the effluent, these parameter concentrations were in the ranges of 0.4–6.3, 0.1–2.4 and 0.01–0.8 g  $1^{-1}$ , respectively. The effluent ranges of BOD<sub>5</sub>/COD ratio was 0.1-0.28. When HRT was decreased from 23 to 12 h, the effluent concentrations of parameters were increased and BOD<sub>5</sub>/COD ratio was decreased. Total COD, SCOD and rbCOD removal efficiencies, respectively, of more than 80. 81 and 90 % were achieved at an OLR of 10.2 g COD  $1^{-1}$  day<sup>-1</sup>, and a gradual decrease in OLR of 25 g COD 1<sup>-1</sup> day<sup>-1</sup> still gave COD, SCOD, rbCOD and BOD<sub>5</sub> removal efficiencies of 75, 63 and 80 %, respectively. Removal efficiency of rbCOD was higher than COD and SCOD. These parameters' removal efficiency was initially high and relatively stable but decreased when OLR was increased to 25 g COD  $1^{-1}$  day<sup>-1</sup>. Average removal efficiency of COD, SCOD and rbCOD in HRT of 24 h was 81, 84 and 91.3 %, respectively. But in HRT of 12 h, it was decreased to 76, 78 and 85 %, respectively.

Some studies revealed good performances of anaerobic sequencing batch reactors. Typical values of 80-90 % and nearly 55 % COD removals were reached in anaerobic laboratory-scale tank at 35 °C and ambient temperature, respectively (Abbas et al. 2009).

#### VFA profile

VFAs effluent concentration was increased to 0.1 g l<sup>-1</sup> when OLR increased to 10.2 g COD  $1^{-1} \text{ day}^{-1}$ . Additionally, when removal efficiency of SCOD was decreased to 77 %, the effluent concentration of VFAs was increased to  $0.34 \text{ g} \text{ l}^{-1}$ . The VFA removal efficiency was high throughout all testing OLRs, in excess of 85 % (Fig. 3). The maximum specific removal rate for VFA-rich wastes such as municipal landfill leachate can be in the range of 4-5 g COD/g VSS d (Hashemi et al. 2015). In Bae et al. study (2010), the level of propionate in the final effluent was lower than 900 mg COD  $1^{-1}$ , indicating relatively good performance of the anaerobic process.





Fig. 3 Effluent mixture VFA concentration in ASBR reactor during the time

#### **Biogas production**

The study was conducted at eleven OLRs from 1.04 to 19.65 g COD  $1^{-1}$  day<sup>-1</sup> in 280 days. The ability of ASBR process in biogas production at various OLRs and its efficiency in removal of COD, SCOD, rbCOD and VFAs were evaluated. At any microorganism concentration, the food-microorganism ratio (F/M) is high right after when the feed cycle is completed. This provides a high driving force for metabolic activity and high overall rates of waste conversion to biogas. Up to the end of the react cycle, the substrate concentration is minimum, providing low F/M ratio for biomass flocculation. The biogas produced during the anaerobic degradation is a valuable resource of energy (Pages-Diaz et al. 2014). The exact expectation of the producible biogas amount and its methane content are one of the most important aspects of an anaerobic reactor. The quality and quantity of the biogas have special importance. The chemical compositions of a leachate determine the potential biogas yields, as well as the gas composition. The biogas production versus OLRs is shown in Fig. 4.

In reactor start-up phase, a lag can be observed in biogas production. Then, during the subsequent batches, the lag becomes less and less pronounced as the reaction becomes quicker. This phenomenon can be attributed to an adaptation of the microorganisms to the waste and to the



Fig. 4 Biogas production at different OLRs in ASBR in the operation times

conditions of the experiment. The biogas produced was 0.55 l when the OLR was 1.04 g COD  $1^{-1} \text{ day}^{-1}$ , and after a gradual increase to an OLR of 10.08 g COD  $1^{-1} \text{ day}^{-1}$ , it still gave 5.68 l.

The range of the methane content of biogas was between 55 and 65 %, which is comparable with the reported range of 67  $\pm$  81 % for the leachate. When the OLR was 1.04 g COD l<sup>-1</sup> day<sup>-1</sup>, the produced methane was 0.3 l, and when gradually increased to an OLR of 10.08 g COD l<sup>-1</sup> day<sup>-1</sup>, it still gave 3.12 l. Methane production for reactor was in the range of 0.29–0.42 l CH<sub>4</sub>/g COD removed. When the OLR was 10.08 g COD l<sup>-1</sup> day<sup>-1</sup> and HRT was 23 h, the biogas production rate was 5.68 l, but in HRT of 12 h and OLR of 19.65 g COD l<sup>-1</sup> day<sup>-1</sup>, the biogas production rate was 6.65 l.

In similar study, about 83 % of COD removed is converted to methane. Also each g of VSS in ASBR reactors is capable of converting a daily maximum of 1.06 g of COD to the methane (Karimi et al. 2011).

### Conclusion

Compost leachate may be treated anaerobically, saving environment and converting the organic material partially to biogas energy.

The COD removal efficiencies were in the range of 76–81 % depending on loading rates applied. The exploitation of produced biogas during anaerobic digestion is an interesting option to achieve a high fuel-to-power conversion rate, even for small-size generators. The maximum volumetric methane production rate (VMPR) of 5.7 1  $CH_4 l^{-1} day^{-1}$  was achieved at the OLR of 19.65 g  $COD l^{-1} day^{-1}$ . About 85 % of COD removed during the treatment was converted to methane. The results have shown that the ASBR reactor could be an appealing option for changing compost leachate into biogas, which may play a serious role in meeting the world's ever-increasing energy requirements in the future.

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